Appendix K Disposal Cell Groundwater Monitoring Plan

DOE/GJ/79491-646

Weldon Spring Site Remedial Action Project

Weldon Spring Site Disposal Cell Groundwater Monitoring Plan

Revision 2

March 2004

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For the

U.S. DEPARTMENT OF ENERGY Grand Junction Operations Office Under Contract DE-AC13-02GJ79491



ABSTRACT

The Weldon Spring Site Disposal Cell Groundwater Monitoring Plan, Rev. 2 describes the approach that will be used to develop tolerance limits on concentrations of contaminants for the cell groundwater monitoring network; the sampling strategy to be implemented for compliance with long-term groundwater monitoring requirements; and outlines the statistical methods to be used in data analysis. The Plan also identifies monitoring well locations, depths and construction details.

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1. INTRODUCTION

1.1 Scope

This plan describes the disposal cell groundwater monitoring program for the U.S. Department of Energy's (DOE's) Weldon Spring Site, which is being conducted according to the substantive requirements of 40 CFR 264, Subpart F, and 10 CSR 25-7.264(2)(F). This plan includes a description of the sampling locations, frequency, parameters, and associated analysis and sampling procedures. A discussion about the data evaluation and the development of the evaluation approach are also included.

1.2 Purpose

The purpose of this plan is to summarize the disposal cell groundwater monitoring program. The following specific elements are addressed: the design of the monitoring network; the results of baseline monitoring; the long-term monitoring program, which includes detection monitoring, compliance monitoring, and corrective action; and data review and reporting.

1.3 Applicable or Relevant and Appropriate Requirements

In the Record of Decision for the Chemical Plant Area of the Weldon Spring Site (Ref. 1), the substantive requirements of 40 CFR 264, Subpart F of the Resource Conservation and Recovery Act (RCRA), and 10 CSR 25-7.264(2)(F) of the Missouri Hazardous Waste Regulations, were identified as applicable or relevant and appropriate requirements (ARARs) for the selected remedy (i.e., construction and operation of an engineered disposal cell). Table 1-1 provides a summary of these ARARs and indicates the sections of this plan that discuss the strategy for meeting each requirement. In addition to these ARARs, relevant portions of 10 CSR 80-3.010(8) were also used as guidance in developing this monitoring plan.

1.4 Background

Groundwater at the chemical plant is contaminated with trichloroethylene (TCE), nitrate, uranium, and nitroaromatic compounds. The groundwater contamination originated with the Raffinate Pits and other source areas of the chemical plant site and former ordnance works area, that have been removed. Contamination is primarily limited to the weathered portion of the uppermost bedrock unit, the Burlington-Keokuk Limestone. Nitroaromatic compounds are present east and north of the disposal cell and is elevated in several of the disposal cell monitoring wells. Nitrate is present north and west of the disposal cell and is elevated in several of the disposal cell monitoring wells. Uranium is present southwest of the disposal cell; however, elevated levels are not observed in any of the disposal cell monitoring wells. TCE is also present southwest of the disposal cell, but elevated levels are not observed in any of the disposal cell monitoring wells.

Table 1-1 ARARs Summary for Disposal Cell Groundwater Monitoring

MOLEN ATION	PERTINENT SECTION OF MONITORING PLAN
40 CFR 264.90 APPLICABILITY Specifies the applicability requirements and exemptions for owners or consistence of facilities that treat store or disnose of hazardous waste.	Section 1.3, Applicable or Relevant and Appropriate Requirements
40 CFR 264.91 REQUIRED PROGRAMS Specifies the criteria for determining which monitoring and response program (i.e., detection monitoring, compliance monitoring, or corrective action) should be instituted at a requiated facility.	Section 4.0, Detection Monitoring Program Section 5.0, Compliance Monitoring and Corrective Action Programs
40 CFR 264.92 GROUNDWATER PROTECTION STANDARD Requires compliance with certain conditions when hazardonis constituents are detected in groundwater underlying a regulated unit.	Section 5.0, Compliance Monitoring and Corrective Action Programs
40 CFR 264.93 HAZARDOUS CONSTITUENTS Specifies the criteria for defining "hazardous constituents" to which the organization standard applies.	Section 5.0, Compliance Monitoring and Corrective Action Programs
40 CFR 264.94 CONCENTRATION LIMITS Specifies the criteria for establishing concentration limits for hazardous constituents detected in the groundwater underlying a regulated unit.	Section 3.3.3, Revised Baseline Tolerance Limits
40 CFR 264.95 POINT OF COMPLIANCE Defines the point of compliance at which the groundwater protection standard applies and monitoring must be conducted.	Section 2.2, Groundwater Monitoring Wells
40 CFR 264.96 COMPLIANCE PERIOD Defines the compliance period during which the groundwater protection standard applies.	Section 5.0, Compliance Monitoring and Corrective Action Programs
40 CFR 264.97 GENERAL GROUNDWATER MONITORING REQUIREMENTS Specifies general requirements for the groundwater monitoring program, such as well installation, sampling and analysis procedures, determination of groundwater surface elevation, and statistical methods to be used.	Section 2.0, Monitoring Network Design Section 3.3.3, Revised Baseline Tolerance Limits Section 4.4, Groundwater Elevation Measurements Section 6.0, Quality Control
40 CFR 264.98 DETECTION MONITORING PROGRAM Specifies requirements for detection monitoring programs, including monitoring parameters, sampling frequency, determination of groundwater flow, determination of statistically significant evidence of contamination, and required response to positive	Section 4.0, Detection Monitoring Program
40 CFR 264.99 COMPLIANCE MONITORING PROGRAM Specifies requirements for compliance monitoring programs, including monitoring parameters, sampling frequency, determination of groundwater flow, determination of statistically significant evidence of contamination, and required response to exceedance of the groundwater protection standard.	Section 5.0, Compliance Monitoring and Corrective Action Programs
40 CFR 264.100 CORRECTIVE ACTION Specifies requirements for corrective actions to be instituted to ansure compliance with the aroundwater protection standard.	Section 5.0, Compliance Monitoring and Corrective Action Programs
10 CSR 25-7.264(2)(F) RELEASES FROM SOLID WASTE MANAGEMENT UNITS Specifies that efforts made to monitor groundwater or implement corrective action be documented, and that daily precipitation be measured. Also requires a surface water monitoring program to represent the quality of surface water hydrologically downgradient of the facility.	Section 2.3, Surface Water Monitoring Location Section 4.5, Precipitation Data Section 4.8, Detection Monitoring Reporting Section 5.0, Compliance Monitoring and Corrective Action Programs

2. MONITORING NETWORK DESIGN

Groundwater monitoring requirements under 40 CFR 264, Subpart F. of the *Resource Conservation and Recovery Act* (RCRA) specify that the monitoring system for a regulated unit must "consist of a sufficient number of wells, installed at appropriate locations and depths, to yield groundwater samples from the uppermost aquifer that: (1) represent the quality of background water that has not been affected by leakage from the regulated unit...; (2) represent the quality of groundwater passing the point of compliance; and (3) allow for the detection of contamination when hazardous waste or hazardous constituents have migrated from the waste management area to the uppermost aquifer." The disposal cell monitoring network at the Weldon Spring Site has been designed to meet these requirements, as described below.

2.1 Basis of Design

The following criteria constitute the basis for design of the disposal cell groundwater monitoring network at the Weldon Spring Site:

- Regulatory requirements,
- Potentiometric surface of the shallow groundwater beneath the disposal cell,
- Design aspects of the disposal cell, and
- Physical site conditions.

The Subpart F regulations of RCRA specify that groundwater monitoring must be conducted at the point of compliance, which consists of a vertical surface that is located hydraulically downgradient of the waste management area and extends down into the uppermost aquifer. The RCRA regulations provide flexibility regarding the number, spacing, and depths of monitoring wells; however, the Missouri Sanitary Landfill regulations in 10 CSR 80-3.010, specify a minimum of one upgradient and three downgradient wells for landfills. The disposal cell network was designed to incorporate one upgradient and four downgradient wells, allowing for the possibility that wells could be added or removed as necessary. Since the original network was installed, two wells have been added and two have been eliminated. Thus, the current network still consists of one upgradient well and four downgradient wells. The location of these wells is discussed in Section 2.2.

To supplement groundwater monitoring, Missouri Hazardous Waste regulations in 10 CSR 25-7.264(2)(F) require that a surface water component be included in monitoring releases from waste management units. The surface water monitoring system must "consist of a sufficient number of points at appropriate locations to yield surface water samples that: (a) represent the quality of background surface water that has not been affected by any contamination from the facility...; and (b) represent the quality of surface water hydrologically downgradient of the facility or regulated units." The surface water monitoring location incorporated in this plan is discussed in Section 2.3.

The potentiometric surface of the shallow groundwater indicates that the flow gradient beneath the disposal cell is generally to the north and northwest, as shown in Figure 2-1. The general direction of groundwater flow has remained relatively unchanged since the cell monitoring system was designed, throughout remediation of the site and construction of the disposal cell. However, since construction of the disposal cell and previous remedial activities, the groundwater elevation has decreased due to dewatering of ponds/basins and diversion of surface water flow and reduced infiltration (recharge) to the shallow aquifer.

Design aspects of the disposal cell that were considered in determining the original locations of the monitoring wells included the locations of the clean fill dikes and leachate collection sump, the 1% to 1.5% northward slope along the base of the cell, and the double liner/leachate collection system. Since the monitoring network was installed while physical site conditions were undergoing frequent change due to remediation and construction activities, existing and planned locations of excavations, roads, structures, surface water bodies, staging areas, and the footprint of the disposal cell were also considered to ensure availability and access to the planned monitoring locations.

2.2 Groundwater Monitoring Wells

The original disposal cell monitoring network was established in 1996. It included five wells: one upgradient well (MW-2048) and four downgradient wells (MW-2032, MW-2045, MW-2046, and MW-2047). The well locations, which are shown in Figure 2-2, were chosen based on the criteria discussed above. Well MW-2048 was installed south of the cell to monitor water quality upgradient of the disposal cell. Wells MW-2045, MW-2046, and MW-2047 were installed northeast, north, and northwest of the cell, respectively, to monitor potential groundwater impacts downgradient of the disposal cell. Well MW-2032 was an existing well that was retained to monitor potential groundwater impacts downgradient (i.e., north) of the leachate sump. Figure 2-3 provides a cross-sectional view of the monitoring system, in relation to the disposal cell and leachate sump.

While the original monitoring network consisted of five wells, it was the intent of the plan to provide flexibility for reacting to the dynamics of the system being monitored. The heterogeneous nature of the fractured bedrock aquifer and the complexities associated with monitoring a previously contaminated groundwater system created uncertainty in the actual performance of the proposed monitoring wells. Additional wells were to be incorporated into the network on an as-needed basis during both the active life and the post-closure period to replace or supplement data from poorly performing wells. Thus, since MW-2045 demonstrated consistently poor hydraulic performance and yielded widely variable analytical data, a fifth downgradient well (MW-2051) was installed in 2001 northeast of the disposal cell, as shown in Figure 2-2. Under the present revision of this plan, MW-2051 replaces MW-2045 as the monitoring location for the northeast side of the disposal cell. Monitoring well MW-2051 exhibits higher hydraulic conductivities and will better represent the shallow groundwater system than MW-2045.

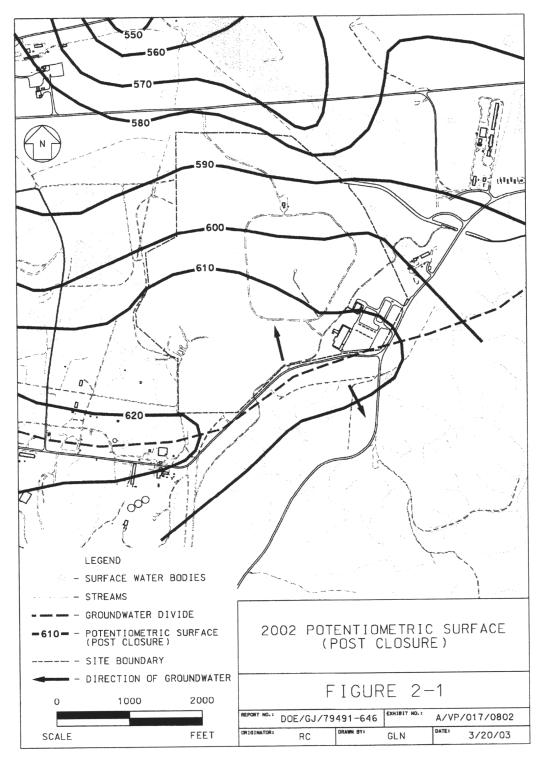


Figure 2-1 2002 Potentiometric Surface (Post Closure)

The original upgradient well, MW-2048 was damaged during construction activities in 2001. This well was determined to be damaged beyond repair, which led to its abandonment and installation of a replacement well shortly thereafter. The new well, MW-2055, is located approximately 20 feet upgradient (i.e., south) of MW-2048 and has replaced it as the upgradient monitoring well (see Figure 2-2). Data review conducted for both monitoring wells indicated comparable upgradient water quality. No PCBs, PAHs or nitroaromatic compounds were detected at either location. Concentration ranges of ions, metals, radiological and indicator parameters were similar, with the exception of three metals (iron, manganese, and nickel) and one indicator parameter (total organic carbon). However, the concentrations are within typical ranges for the groundwater in the weathered Burlington-Keokuk Limestone.

All wells in the disposal cell monitoring network were installed and developed in accordance with 10 CSR 23, *Missouri Water Well Construction Code*. Each well is constructed of 2-inch ID Grade 316 stainless steel casing, with a 10-foot length of 0.010-inch slotted screen. Total depths of the wells range from approximately 45 to 75 feet below ground surface, depending on the respective depth to water at each location. Borehole logs, well diagrams, packer test calculations, and well development forms for the original wells are contained in the *WSSRAP Disposal Cell Monitoring Well Program Installation Report* (Ref. 2). Appendix A of this plan contains the well diagrams, packer test calculations, and well development forms for the two newly installed wells, as well as the borehole logs for all disposal cell wells.

2.3 Surface Water Monitoring Location

The surface water location used to detect downgradient impacts from the disposal cell is Burgermeister Spring (SP-6301) (see Figure 2-4). Historical dye tests have indicated that this spring is the primary localized point of emergence for groundwater from the vicinity of the chemical plant (Ref. 3). Thus, sampling of Burgermeister Spring will yield results that are representative of both surface water and groundwater hydraulically downgradient of the disposal Burgermeister Spring represents the first surface water impacted by groundwater cell. originating from the site, including the disposal cell area. Downstream Lake 34 was not chosen as a monitoring point as Burgermeister Spring represents the worst case conditions for surface water and Lake 34 does not receive surface water contribution from the chemical plant area. It is common practice in aquifer systems dominated by fracture or conduit flow to supplement the monitoring well system by sampling springs that are hydraulically connected to the uppermost aquifer and that have shown a connection to the facility (Ref. 17). This spring has been monitored routinely since 1987 under the Environmental Monitoring Plan (Ref. 4), which contains the overall environmental monitoring requirements for the Weldon Spring site and is a long-term monitoring locations for the Groundwater Operable Unit. There is no upgradient surface water body included in this monitoring plan. The disposal cell is situated near both the regional surface water and groundwater divides; therefore, no surface water bodies are located upgradient of the disposal cell.

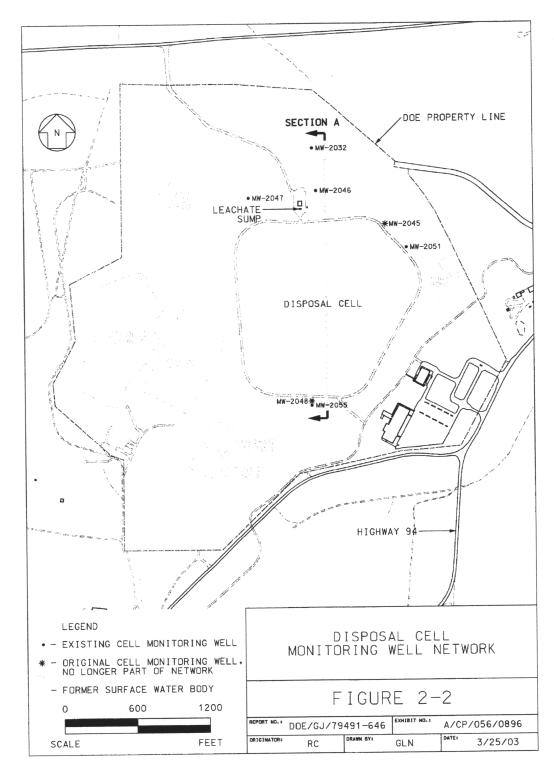


Figure 2-2 Disposal Cell Monitoring Well Network

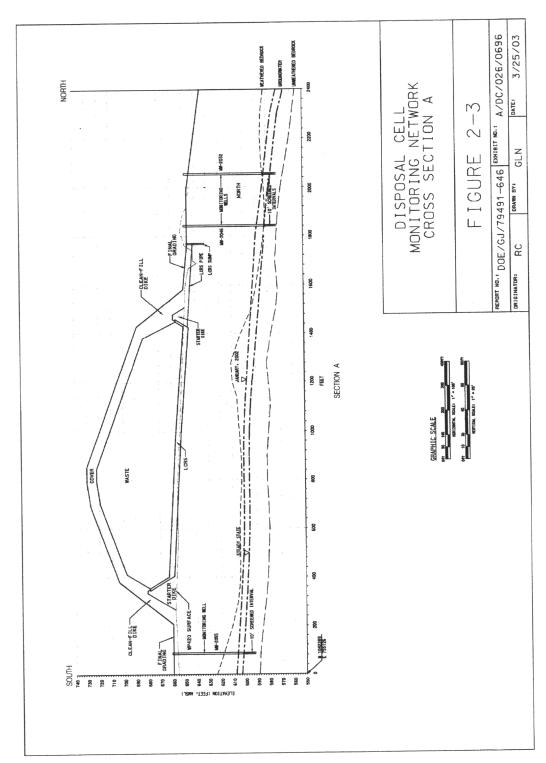


Figure 2-3 Disposal Cell Monitoring Network Cross Section A

Ecological evaluations (including toxicity testing) for Burgermeister Spring have been conducted previously under site environmental monitoring and remedial investigation programs, and these results may be used for a determination of baseline ecological conditions for this plan. Results of ecological studies conducted for Burgermeister Spring as part of the *Remedial Investigation for the Groundwater Operable Unit* (Ref. 3) indicate that current conditions within the surface water and sediments in Burgermeister Spring, while exhibiting above background concentrations of both nitrate and uranium, have not measurably affected the biological community that uses the drainage. Therefore, while sampling for both radiological and chemical constituents will be conducted at Burgermeister Spring as specified in this plan, routine monitoring of biological activity will not be incorporated.

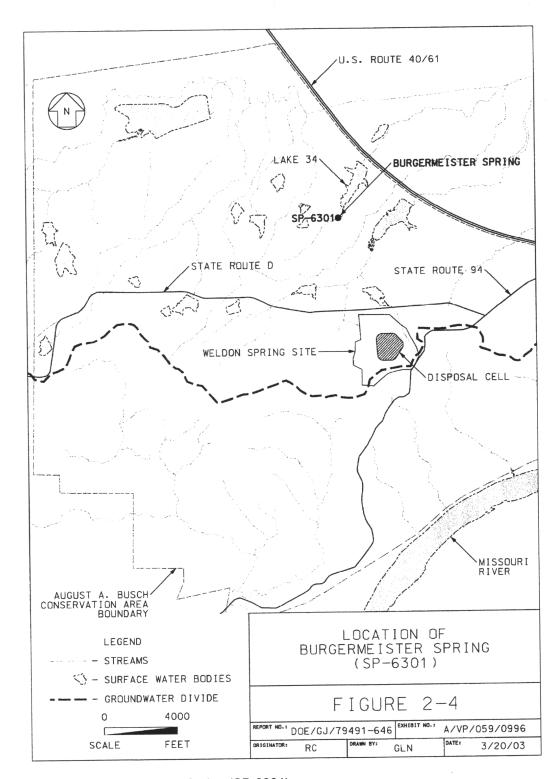


Figure 2-4 Location of Burgermeister Spring (SP-6301)

3. BASELINE MONITORING

In accordance with 40 CFR 264.97, baseline monitoring was conducted to obtain data that represents the quality of groundwater that has not been affected by leakage from the disposal cell. The intent was to establish a baseline data set that could be used in statistical comparisons with detection monitoring data, in accordance with regulatory requirements, to detect and characterize hazardous constituents in the uppermost aquifer that may be due to leakage from the disposal cell.

Review of the initial approach to baseline monitoring and the groundwater system beneath the chemical plant has indicated that in some cases an established baseline may not be appropriate for monitoring of the disposal cell at the Weldon Spring site. The shallow aquifer beneath the chemical plant has been impacted by previous operation of the former ordnance works and the uranium feeds material plant. This limits the reliability of results obtained using the statistical methods specified in the Subpart F regulations to evaluate long term monitoring data, which is discussed further in Section 4.7 and Appendix B. It is expected that groundwater conditions for the contaminants of concern for the Groundwater Operable Unit (nitrate, uranium, trichloroethylene, and nitroaromatic compounds) will improve over time due to source removal activities. Baseline values established for these contaminants using prior data may be biased high due to existing groundwater contamination or contamination resulting from contaminated soil remediation. Later comparisons to this baseline may mask trends in the groundwater.

3.1 Initial Baseline Monitoring

Baseline monitoring of locations MW-2032, MW-2045 through MW-2048, and SP-6301 was conducted throughout 1997 and early 1998, prior to waste placement activities. Four replicate samples were obtained from each location on a quarterly basis for approximately one year (i.e., five separate sampling events resulting in twenty individual samples per well). The initial baseline data indicated a large degree of temporal and spatial variability in water quality at the monitoring locations, as evidenced by the wide range of mean concentrations among monitoring locations and the high standard deviations calculated for many of the parameters. This variability is due to several contributing factors, such as the heterogeneity of the naturally occurring geochemistry, the variations in historical contaminant distribution throughout the site, and the unpredictable flow patterns in the fractured bedrock environment. Thus, the baseline conditions represented by the data were actually an indication of the groundwater quality at a particular location over a particular time period, and not a definitive characterization of background as intended by the Subpart F regulations.

Baseline monitoring of wells MW-2051 and MW-2055 began in 2001 and 2002, respectively. Samples from these wells are considered to be representative of water quality not impacted by the disposal cell since previous groundwater and leachate monitoring have indicated no reason to suspect leakage from the cell. Elements of the baseline sampling at these two wells

have been similar to those listed above for the original six locations, except that the sampling events have consisted of a single sample (i.e., no replicates other than for QC purposes).

3.2 Previous Leachate Monitoring Evaluation

Regulations contained in 40 CFR 264.301 require leachate to be monitored during the active operation and post-closure period of a hazardous waste landfill. Although not specifically addressed by groundwater monitoring regulations, leachate monitoring is discussed in this plan because of the need to correlate the two programs to effectively monitor the potential migration of contaminants from the disposal cell.

Leachate production and analytical data have been collected routinely since waste placement activities began, in accordance with the *Disposal Cell Leachate Monitoring Plan* (Ref. 5) and the *Long-Term Stewardship Plan for the Weldon Spring, Missouri, Site* (Ref. 6). Samples were collected at least quarterly and analyzed for the entire list of parameters shown in Table 3-1. A summary of the average and maximum concentrations of analytical constituents detected in the leachate since the sump was completed (2000) is also presented in Table 3-1.

Table 3-1 Leachate Monitoring Data (10/18/00 to 5/8/02)

Dto		Concentration in Leachate (10-1		
Parameter	Units	Average	Maximum	
Chloride ^(b)	(mg/l)	30.40	38.80	
Fluoride ^(b)	(mg/l)	0.24 0.29		
Nitrate-N ^(b)	(mg/l)	0.56	3.10	
Sulfate ^(b)	(mg/l)	94.63	163.00	
Aluminum	(μ g/l)	33.12	70.50	
Antimony	(μg/l)	ND	ND	
Arsenic ^(b)	(μg/l)	3.73	9.36	
Barium ^(b)	(μg/l)	606.88	832.00	
Beryllium	(μg/l)	0.41	0.92	
Cadmium	(μ g /l)	ND	ND	
Calcium	(mg/l)	176.25	198.00	
Chromium ^(b)	(μ g/l)	ND	ND	
Cobalt ^(b)	(μg/l)	17.23	25.90	
Copper	(μg/l)	3.48	9.90	
Iron ^(b)	(μg/l)	12,083.00	22,100.00	
Lead ^(b)	(μg/l)	ND	ND	
Lithium	(μg/l)	7.99	13.20	
Magnesium	(mg/l)	52.41	55.70	
Manganese ^(b)	(μ g /l)	5,396.00 9,97		
Mercury	(μ g/ l)	ND	ND	
Molybdenum	(μg/l)	5.82 7.75		
Nickel ^(b)	(μ g /l)	9.71	14.70	
Potassium	(mg/l)	5.40	6.29	

Table 3-2 Leachate Monitoring Data (10/18/00 to 5/8/02) (Continued)

Danamatan IInita		Concentration in Leachate (10-18-2000 to 5-8-		
Parameter	Units	Average	Maximum	
Selenium ^(b)	(μ g/l)	1.24	3.95	
Silver	(μ g /l)	ND	ND	
Sodium	(mg/l)	69.49 77.		
Thallium ^(b)	(μ g/l)	3.45	10.60	
Vanadium	(μ g/l)	0.99	2.00	
Zinc	(μ g/l)	22.76	40.90	
C.O.D. ^(b)	(mg/l)	28.60	35.00	
Cyanide	(μ g/l)	2.91	6.10	
T.D.S. ^(b)	(mg/l)	867.20	970.00	
T.O.C. ^(b)	(mg/l)	9.42	10.50	
1,3,5-TNB ^(b)	(μg/l)	ND	ND	
1,3-DNB ^(b)	(μg/l)	ND	ND	
2,4,6-TNT ^(b)	(μ g/l)	ND	ND	
2,4-DNT ^(b)	(μg/l)	ND	ND	
2,6-DNT ^(b)	(μ g/l)	ND		
Nitrobenzene ^(b)	(μ g/l)	ND	ND	
Gross alpha	(pCi/l)	66.44	180.00	
Gross beta	(pCi/l)	28.56	59.60	
Ra-226 ^(b)	(pCi/l)	0.32	0.68	
Ra-228 ^(b)	(pCi/l)	0.60	1.37	
Th-228 ^(b)	(pCi/l)	0.10	0.34	
Th-230 ^(b)	(pCi/l)	0.23	0.36	
Th-232 ^(b)	(pCi/l)	0.09	0.25	
Total Uranium ^(b)	(pCi/l)	75.54	278.00	
Pesticides	(μ g/ l)	ND ^(a)	0.26	
PCBs ^(b)	(μg/l)	ND	ND	
PAHs ^(b)	(μ g/l)	ND	ND	
VOCs	(μ g/ l)	ND ^(a)	5.20	

Notes:

ND = non-detect

3.3 Evaluation of Baseline Data

The original disposal cell monitoring plan specified that groundwater and surface water samples obtained under the plan be analyzed for all constituents presented in Table 3-2. This comprehensive list included general water quality indicator parameters (e.g., pH, temperature, total organic carbon, etc.), chemical and radiological contaminants, and naturally occurring constituents. The list included many parameters in addition to those that would be considered "hazardous constituents" under 40 CFR 264.93, and provided the basis for a thorough assessment of groundwater quality in the vicinity of the cell.

⁽a) All data were reported as non-detect, except for 3 or 4 isolated detections of individual compounds.

(b) These parameters are retained for leachate analysis as of the date of this plan.

It was anticipated that the original list of analytical parameters would be evaluated periodically and modified as necessary to eliminate constituents that could provide no conclusive information regarding the presence of hazardous constituents due to a potential breach in the cell liner system. The first such modification was instituted in December 1999, after the initial baseline data had been evaluated and the detection monitoring program had begun (Ref. 7). Several parameters were eliminated from the list due to the lack of measurable detections in either the groundwater or the leachate, or because they were naturally occurring parameters that were not site contaminants of concern (see footnote "a" on Table 3-1).

Table 3-3 Constituents Monitored in Groundwater and Surface Water

General			Contaminants		
Indicator Parameters	Metals	Nitroaromatic Compounds	Radiological	Inorganic Ions	Other
pH ^(b) Temperature ^(b) Specific Conductance ^(b) COD ^(b) Cyanide ^(c) TDS ^(b) TOC ^(b) TOX ^(c)	Aluminum ^(c) Antimony ^(c) Arsenic ^(b) Barium ^(b) Beryllium ^(a) Cadmium ^(a) Calcium ^(a) Chromium ^(b) Copper ^(c) Iron ^(a) (b) Lead ^(b) Lithium ^(c) Magnesium ^(c) Magnesium ^(c) Molybdenum ^(c) Nickel ^(b) Potassium ^(a) Selenium ^(b) Silver ^(c) Sodium ^(a) Thallium ^(a) (b) Vanadium ^(c) Zinc ^(c)	1,3,5-TNB ^(b) 1,3-DNB ^(b) 2,4,6-TNT ^(b) 2,4-DNT ^(b) 2,6-DNT ^(b) Nitrobenzene ^{(a)(b)}	Radium-226 ^(b) Radium-228 ^(b) Thorium, Isotopic ^(b) Uranium, Total ^(b)	Chloride ^(b) Fluoride ^(b) Nitrate-N ^(b) Sulfate ^(b)	PCBs ^{(a)(b)} PAHs ^{(a)(b)} VOCs ^(a) Pesticides ^(c)

COD Chemical Oxygen Demand **Total Dissolved Solids** TDS

Polychlorinated Biphenyls: Aroclor 1248, 1254, 1260 **PCBs PAHs**

Polyaromatic Hydrocarbons: benz(a)anthracene, benzo(b)fluorancene, benzo(k)fluorancene, benzo(a)pyrene,

Total Organic Carbon TOC **VOCs**

Volatile Organic Compounds Total Organic Halogen

chrysene. indeno(1,2,3-cd)pyrene

TOX (a) These parameters were deleted from the list in December 1999 because either they had not been detected previously in any measurable quantities or they were naturally occurring parameters that were not contaminants of concern (Ref. 7).

⁽b) These parameters are retained or reinstated for groundwater and surface water analysis as of the date of this plan.

⁽c) These parameters are eliminated as of the date of this plan.

3.3.1 Identification of Long-Term Monitoring Parameters

This revision further modifies the list of groundwater monitoring parameters based on a review of the Chemical Plant and Quarry Bulk Waste Operable Units contaminants of concern, materials known to be present in the disposal cell waste, and leachate analytical data. The following contaminants of concern were identified in wastes from the chemical plant and/or the arsenic, chromium, lead, nickel, selenium, thallium, nitroaromatic quarry bulk waste: compounds (specifically 2,4,6-DNT), radium, thorium, uranium, polychlorinated biphenyls (PCBs), and polyaromatic hydrocarbons (PAHs) (Refs. 1 and 8). In addition, barium, manganese, and selenium were determined to be present in the water treatment processing wastes during the remediation of contaminated surface water. As leachate analytical data have become available, the following constituents have been identified as being present at relatively higher concentrations in the leachate than in the underlying groundwater: arsenic, barium, cobalt, iron, manganese, uranium, and COD. These parameters are important to the cell monitoring network because a breach of the cell liner system could result in detectable increases in the levels of these constituents in the groundwater.

Since the above contaminants are known to be present in the disposal cell waste, it is possible that they may become constituents in the cell leachate and, if there is a breach in the system, eventually be detected in the underlying groundwater. Thus, the above contaminants are identified as monitoring parameters for this program.

3.3.2 Identification of Signature Parameters

Detection monitoring data obtained from the cell well network from 1998 through 2001 were evaluated using several of the suggested statistical methods in an attempt to identify statistically significant evidence of contamination due to the disposal cell. Results of these evaluations, which are summarized in Appendix B, demonstrate the uncertainties associated with applying the prescribed methods to data from an aquifer with preexisting contamination and where a high degree of spatial variation in contaminant distribution exists among the monitoring wells. Each type of evaluation resulted in numerous "false positive" statistical failures that, rather than providing reliable and conclusive evidence of cell leakage, were attributable to fluctuations in preexisting groundwater contamination.

The list of monitoring parameters in Table 3-2 includes indicator parameters and waste constituents that for an uncontaminated aquifer would provide a reliable indication of the presence of hazardous constituents in groundwater due to leakage from the disposal cell. However, most of these parameters are already present in the groundwater at higher levels than in the leachate, either naturally or due to historical contamination, or are not present in either the groundwater or the leachate at concentrations above the detection limit. Thus, most of the parameters on this list are not able to provide conclusive evidence of cell leakage since impacts from the leachate would not cause detectable changes in the underlying groundwater.

The most reliable means of detecting potential impacts due to leakage of the disposal cell is to focus on parameters that exist at significantly higher concentrations in the leachate than in the groundwater. An increasing trend in these parameters in the groundwater would be detectable and, most likely, attributable to cell leachate since all other sources have been remediated

To this end, the following constituents have been identified as "signature parameters" for the disposal cell detection monitoring program: barium, uranium, iron, and manganese. All four of these parameters have been detected at concentrations at least an order of magnitude higher in the leachate (Table 3-1) than in the underlying groundwater or Burgermeister Spring (with the exception of uranium), which enhances the reliability of any conclusions that are drawn based on fluctuations in groundwater constituents. Increasing trends of these four parameters in the groundwater will be considered a signature of cell leachate that has migrated to the underlying aquifer and additional actions will be taken as described in Section 4.7. also, these four parameters are naturally occurring and with the exception of uranium should not change via attenuation overtime. Uranium, a contaminant of concern for the Groundwater Operable Unit, is expected to attenuate with time where uranium impact occurs. However, the activity measured in the disposal cell monitoring wells is similar to background and likely will not change substantially over time. It is anticipated that the list of signature parameters may be modified, as necessary, based on future changes in leachate and/or groundwater concentrations.

It should be noted that the uranium concentrations in Burgermeister Spring can be similar or higher than those exhibited in the leachate. This location is impacted by not only contaminated groundwater originating from the Raffinate Pit area, but also residual contamination that is present in the losing stream segment that extends from the Ash Pond area of the site to Burgermeister Spring. Increasing trends in uranium should not be used as the only indicator of possible leakage from the disposal cell.

3.4 Statistical Analysis of Data

3.4.1 Distribution of Data

The data for the signature parameters at the cell wells locations were examined to determine whether the data is normal or log-normal (Appendix B). The data shows a stronger evidence of log-normality than normality. However, to demonstrate that there is little difference in the method used to calculate the baseline tolerance limits, values were calculated for the signature parameters at three of the locations using six methods. The methods used were: EPA guidance suggested method on normal and log-normal data, tolerance limits on normal and log-normal data, and the mean plus 3 standard deviations on normal and log-normal data. All of the data from each location was used in this evaluation. The values calculated using the six methods yielded similar values for each of the signature parameters. Based on the evaluation (Appendix B), it is recommended to maintain the existing methodology of calculating baseline tolerance limits for the signature parameters and assume the data is distributed normally. Every 5 years,

likely in conjunction with the CERCLA five-year reviews, the distribution of the data will be reevaluated.

3.4.2 Revised Baseline Tolerance Limits

Tolerance limits for signature parameters have been calculated using the dataset from 1997 through 2002, using 95% confidence and 95% coverage, based on the assumption that the data are normally distributed (Table 3-3). In the case of the newer wells (MW-2051 and MW-2055), the available data used is fairly small: however the tolerance limits for these wells are representative of groundwater conditions at these locations. Every 5 years, likely in conjunction with the CERCLA five-year reviews, the baseline tolerance limits will be recalculated.

Table 3-3 Baseline Tolerance Limits for Signature Parameters in Groundwater and Surface Water	

	Signature Parameter				
Location	Barium (μg/l)	Iron (μg/l)	Manganese (μg/l)	Uranium (pCi/l)	
MW-2032	377	1,125	57	6.4	
MW-2046	277	1,578	187	1.8	
MW-2047	471	1,485	171	2.7	
MW-2051	285	2,896	265	4.5	
MW-2055	98	10,579	179	7.5	
SP-6301	180	2,608	88	159	

In calculating these values, results reported as non-detect (ND) or less than the detection limit (DL) were assigned a value of one-half the DL. Estimated values less than the detection limit, when reported, were used rather than one-half the DL.

In accordance with the U. S. EPA guidance on *Statistical Analysis of Groundwater Monitoring Data at RCRA Facilities* (Ref. 10), the following formula was used to calculate baseline tolerance limits (BTLs):

$$BTL = x + k(s)$$

where: x = arithmetic mean of the baseline data

s = standard deviation of the baseline data

k = one-sided normal tolerance factor, based on number of values in the data set

The tolerance limits for each location using data collected through December 2002 is included in Appendix B. One-sided tolerance factors can be found in Table 5 - Appendix B of the EPA guidance (Ref. 10) and are also included in Appendix B.

4. DETECTION MONITORING PROGRAM

The goal of the detection monitoring program is to be able to detect releases of hazardous constituents from the disposal cell to the underlying aquifer. Detection monitoring is conducted in accordance with 40 CFR 264.98 throughout the life of the disposal cell to allow for the detection of hazardous constituents that may be migrating from the disposal cell.

The detection monitoring program, which began at this site in June 1998, has evolved since its inception as additional groundwater and leachate data have been obtained and evaluated in light of the relevant regulatory requirements. Resulting modifications to the plan have been incorporated through correspondence (Ref. 7 and 11), annual revisions to the site *Environmental Monitoring Plan* (Ref. 4), and this revision of the *Weldon Spring Site Disposal Cell Groundwater Monitoring Plan*.

4.1 Sampling Locations

Samples will be collected from monitoring wells MW-2032, MW-2046, MW-2047, MW-2051, and MW-2055. Samples will also be collected from Burgermeister Spring (SP-6301).

4.2 Parameters

Samples collected from the monitoring wells and Burgermeister Spring will be analyzed for the list of parameters given in Table 4-1. Quality control sampling is discussed in Section 7.

Table 4-1 Detection Monitoring Parameter List for Groundwater and Surface Water				
D- diele vieel	Inorganic	Motale	Nitroaromatic	Other

Radiological	Inorganic Ions	Metals	Nitroaromatic Compounds	Other	General Indicator Parameters
Radium-226 Radium-228 Thorium, Isotopic Uranium, Total *	Chloride Fluoride Nitrate (as N) Sulfate	Arsenic Barium * Chromium Cobalt Iron * Lead Manganese * Nickel Selenium Thallium	1,3,5-TNB 1,3-DNB 2,4,6-TNT 2,4-DNT 2,6-DNT Nitrobenzene	PCBs PAHs	pH Temperature Specific Conductance COD TDS TOC Turbidity

Signature parameters (see Section 3.3.2)

4.3 Sampling Frequency

Each monitoring well and Burgermeister Spring will be sampled on a semiannual frequency. Samples will be collected during June and December of each year. This sampling

frequency will provide an adequate dataset for use in developing a moving baseline for each location (Section 3.3), and assists in eliminating the spatial and temporal variability seen in earlier datasets. Burgermeister Spring will be sampled during baseflow conditions, which is the stage of spring discharge when the water is least influenced by active surface runoff. Samples will be collected no sooner than 1 week following the end of a precipitation event of sufficient intensity to result in surface runoff. The flow rate of the spring will be estimated and recorded at each sampling event.

The original disposal cell groundwater monitoring plan called for collecting four replicates at each monitoring location on a semi-annual basis. In 1999, the monitoring frequency was reduced to a single sample collected semi-annually from each location since independent replicates could not be collected within a short time period because of slow groundwater flow rates.

4.4 Groundwater Elevation Measurements

Groundwater elevations will be measured semiannually at each of the disposal cell monitoring well locations prior to each sampling event. Results for 1997 through 2002 are presented in Figure 4-1. Groundwater elevations have remained relatively constant since the wells were installed. Groundwater flow rates and flow directions will be evaluated annually. A presentation of the potentiometric surface and determination of the flow rates and directions for 1998 through 2002 are presented in Appendix C.

4.5 Precipitation Data

To support leachate monitoring activities at a regulated unit, Missouri Hazardous Waste regulations require the collection of local precipitation data. An onsite meteorological station was used to monitor daily and hourly precipitation until December 2001, as described in the *Environmental Monitoring Plan* (Ref. 4). More recent and future regional precipitation data (e.g., from the Spirit of St. Louis Airport in Chesterfield, MO) is obtained as needed through the National Oceanographic and Atmospheric Administration at the following internet address:

4.6 Leachate Monitoring

Regulations contained in 40 CFR 264.301 require leachate to be monitored during the active operation and post-closure period of a hazardous waste landfill. Although not specifically addressed by groundwater monitoring regulations, leachate monitoring is discussed in this plan because of the need to correlate the two programs to effectively monitor the potential migration of contaminants from the disposal cell.

This plan revision modifies the leachate monitoring parameters to be the same as the list of parameters monitored in the groundwater. The leachate will continue to be monitored

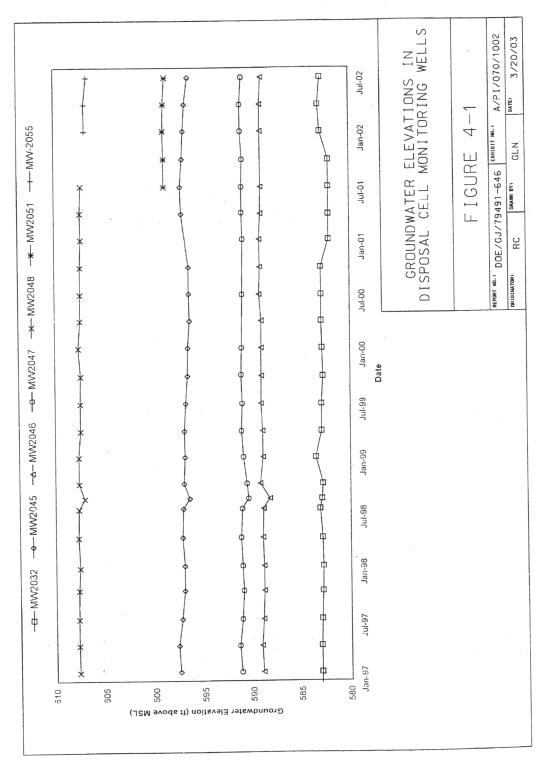


Figure 4-1 Groundwater Elevations in Disposal Cell Monitoring Wells

semiannually for the parameters outlined in Table 4-1. Samples will be collected in June and December of each year.

4.7 Detection Monitoring Data Review

4.7.1 Signature Parameters

Under the detection monitoring program, data for only the signature parameters from each monitoring event will be compared to baseline tolerance limits (Table 3-3) to track general changes in groundwater quality and determine whether statistically significant increases in these parameters has occurred. Previously, all the parameters were evaluated against the appropriate baseline tolerance limits and exceedances were attributable to variations in existing groundwater quality, interference from turbid samples, leaching of metals (chromium, nickel, cobalt, and molybdenum) from stainless steel (Type 316) well materials, sample preservation or analytical error, or inherent uncertainties in data that are less than five times greater than the detection limit.

For signature parameters (barium, iron, manganese, and uranium) that are determined to exceed the baseline tolerance limits, the following actions will be taken:

- The location will be resampled to confirm the exceedence. If the exceedence is not confirmed, detection monitoring will continue and no further action is necessary.
- If resampling results confirm the exceedence, a thorough evaluation will be performed to determine whether it is due to leakage from the disposal cell. This evaluation may include an assessment of groundwater gradients, review of leachate production and analytical data, review of sitewide monitoring data, and additional sampling. If it is shown that the upward trend is not due to leakage from the cell, a demonstration report will be prepared in accordance with the substantive requirements of 40 CFR 264.98, and detection monitoring will continue.

4.7.2 Other Parameters

The data from the remainder of the parameters will be reviewed to evaluate the general groundwater quality in the vicinity of the disposal cell and to determine if changes are occurring in the groundwater system. Data will be compared to the 3 most recent years of data to determine if statistically significant increases or trends in concentrations are present. A "moving average" approach, as discussed in the October 11, 1989 Federal Register (Ref, 9), is used to better reflect naturally occurring changes in site hydrogeology, minimize temporal variations, and account for the natural attenuation of contaminants in the shallow aquifer. Data will be considered statistically significant if it is greater than the arithmetic mean plus 3 times the standard deviation for each location.

Data that are determined to be statistically significant will be evaluated as follows:

- The location will be resampled to confirm the exceedence. If the exceedence is not confirmed, no further action is necessary.
- If results of the resampling confirm the exceedence, the data will be compared to the leachate data. If the leachate data do not indicate that the exceedence could be the result of leakage from the cell (parameter is not elevated in the leachate), an assessment of the analytical data and review of sitewide monitoring data will be performed. If the exceeding parameter is a contaminant of concern for the Groundwater Operable Unit (nitrate, nitroaromatic compounds, or trichloroethylene), this information will be evaluated under the monitoring program for the Groundwater Operable Unit at the chemical plant.
- If results of the resampling confirm the exceedence, the data will be compared to the leachate data. If the leachate data indicate that the exceedence could be the result of leakage from the cell (parameter is also elevated in the leachate), the entire disposal cell monitoring network will be sampled for the full list of parameters shown in Table 3-2. A revised monitoring plan, which incorporates the results of the enhanced sampling and outlines the specific details of the compliance monitoring program (Section 5), and an engineering feasibility plan for corrective action will be prepared in accordance with substantive requirements of 40 CFR 264.99.

4.7.3 Leachate

Analytical data from the leachate will be compared to the analytical data from the monitoring well network to determine the adequacy of the signature parameters for this plan. If the composition of the leachate changes substantially, a parameter may be included or removed from the signature parameter list. If the concentration of a parameter decreases so that it is not distinguishable from concentrations (similar in concentrations) in groundwater, that parameter will be removed from the signature parameter list. Conversely, if the concentration of a parameter increases to a level that distinguishable from the concentrations in groundwater (order of magnitude greater), it would warrant its inclusion in the signature parameter list. This evaluation will be performed annually.

4.8 Detection Monitoring Reporting

4.8.1 Annual Reporting

Disposal cell monitoring data are reported annually in the *Weldon Spring Site Environmental Reports* (Refs. 12, 13, 14, and 15). Data to be reported includes all detectable analytical results, as well as groundwater flow rate and direction. However, since only analytical results were reported prior to 2002, groundwater flow rates and direction for the years 1998 through 2002 are included in Appendix C of this plan.

Confirmed exceedances of signature parameters are investigated further by evaluating water levels and precipitation data and reviewing historical analytical and field monitoring data to determine the likely cause and contributing factors. A summary of the exceedances and results of the investigations are reported both in the annual site environmental report and in the quarterly Federal Facilities Agreement (FFA) Report.

4.8.2 Demonstration Reporting

A demonstration report will be prepared if it is shown that an upward trend in a signature parameter is not due to leakage from the cell. This report will document the evaluation used to derive the conclusion that leakage has not occurred from the disposal cell. This evaluation may include an assessment of data quality, groundwater gradients, review of leachate production and analytical data, review of sitewide monitoring data, and additional sampling.

5. COMPLIANCE MONITORING AND CORRECTIVE ACTION PROGRAMS

If it is determined that leakage from the cell has resulted in deterioration of the groundwater at the chemical plant, a review of the remedy will be necessary. This is based on the condition that the remedy is not behaving as expected and may no longer be protective of human health and the environment. Modifications or actions would be documented under CERCLA and would be consistent with 40 CFR 264.100, if appropriate. Identification of applicable or relevant and appropriate regulations (ARARs) would be make at that time and my include groundwater protection standards as outlined in 40 CFR 264.92, if appropriate. At that time, a modification of this program would be documented in collaboration with the U.S. Environmental Protection Agency – Region VII and the Missouri Department of Natural Resources. The monitoring program will continue as prescribed in Section 4 and the nature and extent of the release will be investigated.

6. QUALITY CONTROL

6.1 Sampling and Analysis Procedures

The general groundwater monitoring requirements in 40 CFR 264.97 specify that the monitoring program for a regulated unit must incorporate consistent, reliable, appropriate, and accurate sampling and analysis procedures. The *Ground Water and Surface Water Sampling and Analysis Plan for GJO Projects* (Ref. 16) establishes the data quality requirements for all environmental data collected at the Weldon Spring Site, including data obtained in support of the disposal cell groundwater monitoring plan. Standard operating procedures (SOPs) have been developed and implemented to provide consistency in sample collection methodology and documentation of environmental activities. The appropriate sections of the *Sampling and Analysis Plan* identified below apply to all monitoring activities conducted under this plan.

6.1.1 Field Documentation

Water elevations, sample locations, water temperatures, and other physical parameters are recorded in the field. This information will be recorded and documented as provided in Section 4.3 of the Sampling and Analysis Plan.

6.1.2 Field Measurements and Equipment Calibration

Prior to sample collection, specific field parameters are measured. These include the physical parameters listed above and groundwater elevations. Procedures for obtaining these measurements and calibration of equipment are provided in Section 3.6 of the *Sampling and Analysis Plan*.

6.1.3 Sample Identification

All samples, including quality control samples, collected under this monitoring plan are identified with a unique sample identification number, according to Section 6.0 of the *Sampling and Analysis Plan*.

6.1.4 Sample Collection, Preparation, and Preservation

Procedures for collecting groundwater and spring samples are defined in Section 3.7 of the *Sampling and Analysis Plan*. All samples collected under this plan will be unfiltered. Table 7-1 lists the general sample preparation and preservation requirements for each parameter. All samples taken are collected in certified-clean plastic, clear glass, or amber glass bottles, as appropriate for analysis. Sample packaging and shipping conforms to Section 6.0 of the *Sampling and Analysis Plan*.

Table 6-1 Sample Preparation and Preservation Requirements

ANALYSIS	SAMPLE CONTAINER SIZE/TYPE	PRESERVATION	HOLDING TIME ^(a)	VOLUME REQUIRED (Minimum)	MS/MD OR DU VOLUME REQUIRED
Nitroaromatic Compounds	1-liter amber glass	4°C (ice)	7 days	1000 ml	3000 ml
Metals	1-liter plastic	HNO ₃ - pH of <2	6 months (Hg: 28 days)	500 ml	1000 ml
Sulfate Fluoride Chloride TDS	1-liter plastic	4°C (ice)	28 days (TDS-7 days)	300 ml	1000 ml
Uranium, total Thorium, isotopic Radium-226 Radium-228	4-liter plastic cubit	HNO ₃ - pH of <2	6 months	4 liters	12 liters
PCBs	1-liter amber glass	4°C (ice)	7 days	1000 ml	3000 ml
PAHs	1-liter amber glass	4°C (ice)	7 days	1000 ml	3000 ml
Nitrate (as N) TOC TOX COD	500-ml amber glass	H₂SO₄ - pH of <2	28 days	300 ml	1000 ml
VOCs	40-ml vial	4°C (ice) HCl - pH of <2	14 days	80 ml (2 vials)	160 ml (4 vials)

MS Matrix Spike

Matrix Spike Duplicate

DU Duplicate HCI – Hydrochloric Acid

 HNO_3 – Nitric Acid H_2SO_4 – Sulfuric Acid HCI – Hydrochloric Acid (a) Actual extraction/analysis holding times are variable. Samples should be shipped immediately after collection.

6.1.5 Chain-of-Custody

Chain-of-Custody forms are maintained for all environmental samples collected. The chain-of-custody process is detailed in Section 6.0 of the *Sampling and Analysis Plan*. This section also outlines specific instructions for ensuring that samples are not tampered with or altered prior to analysis.

6.1.6 Sampling Equipment Decontamination

All groundwater wells have dedicated bladder pumps and hoses. Other sampling equipment is decontaminated as necessary according to Section 7.0 of the *Sampling and Analysis Plan*.

6.1.7 Analytical Procedures

Analytical testing is conducted by either the GJO Analytical Chemistry Lab or by subcontracted laboratories (nitroaromatic compounds) that follow the EPA Contract Laboratory Program (CLP) requirements for metals and organic compounds, the EPA drinking water and

radiochemical methodologies for other parameters, or alternate methods, as described in Section 5.0 of the Sampling and Analysis Plan.

Detection limits are specified in contracts established with the laboratories. In general, these detection limits follow CLP protocol and standard analytical methodology. Table 7-2 provides the detection limits and analytical methods used for the disposal cell groundwater monitoring program.

Table 6-2 Specified Detection Limits and Analytical Methods

Analytical Parameter	Analytical Method	Required Detection Limit	
WET CHEMISTRY PARAMETERS			
Total Organic Halides	SW-846 9020A	0.5 mg/l	
Chemical Oxygen Demand	EPA 410	5.0 mg/l	
Total Organic Carbon	EPA 415.1	0.1 mg/l	
Total Dissolved Solids	EPA 160.2	1.0 mg/l	
RADIOLOGICAL CONSTITUENTS			
Total Uranium	ASTM 5174-91 (Fluorimetry or KPA) or equivalent	1.0 pCi/l	
Isotopic Thorium	ASTM, EPA, EML or equivalent	0.2 pCi/l (each isotope)	
Radium-226	ASTM, EPA, EML or equivalent	1.0 pCi/l	
Radium-228	ASTM, EPA, EML or equivalent	5.0 pCi/l	
NITROAROMATIC COMPOUNDS			
2,4-DNT	USATHAMA or EPA 8330	0.030 μg/l	
2,6-DNT	USATHAMA or EPA 8330	0.010 μg/l	
2,4,6-TNT	USATHAMA or EPA 8330	0.030 μ g/ l	
1,3,5-TNB	USATHAMA or EPA 8330	0.030 μg/l	
1,3-DNB	USATHAMA or EPA 8330	0.090 μg/l	
Nitrobenzene	USATHAMA or EPA 8330	0.030 μg/l	
INORGANIC IONS			
Nitrate-N	EPA 300/340	0.10 mg/l	
Sulfate	EPA 300/375	2.0 mg/l	
Fluoride	EPA 300/375	0.1 mg/l	
Chloride	EPA 300/375	0.25 mg/l	
METALS			
All	Contract Lab Program	Instrument Detection Limits (IDLs)	
ADDITIONAL CONSTITUENTS			
PAHs	SW846 8310	5.0 μg/l (each parameter)	
PCBs	EPA 608 or SW846 8081a/8082	1.0 μg/l (each parameter)	

6.2 Quality Control Samples

Quality control (QC) samples are collected to ensure consistent and accurate performance of sample collection and laboratory analysis activities. Section 4.0 of the *Sampling and Analysis Plan* defines the QC samples to be collected, the recommended collection frequency, and the collection procedures. Table 7-3 lists the types of quality control samples that will be collected under this plan and identifies their purpose in support of the monitoring program.

Table 6-3 Field Quality Control Sample Summary

QUALITY CONTROL SAMPLE TYPE	FREQUENCY	PURPOSE
Duplicate/Matrix Spike/Matrix Spike Duplicate	1 per 20 samples	Assess laboratory method variability.
Field Replicate	1 per 20 samples	Assess matrix variability.
Deionized Water Blank	1 per month	Assess quality of deionized water used to decontaminate water level meter.

6.3 Data Review

The Sampling and Analysis Plan (Section 5 and Appendix B) describes the verification, validation, and technical review process to which data obtained under this plan are subject. Analytical data obtained under this plan are maintained in the GJO SEEPro database. This database allows for data input, storage, and retrieval so that the statistical analyses required under this plan can be performed.

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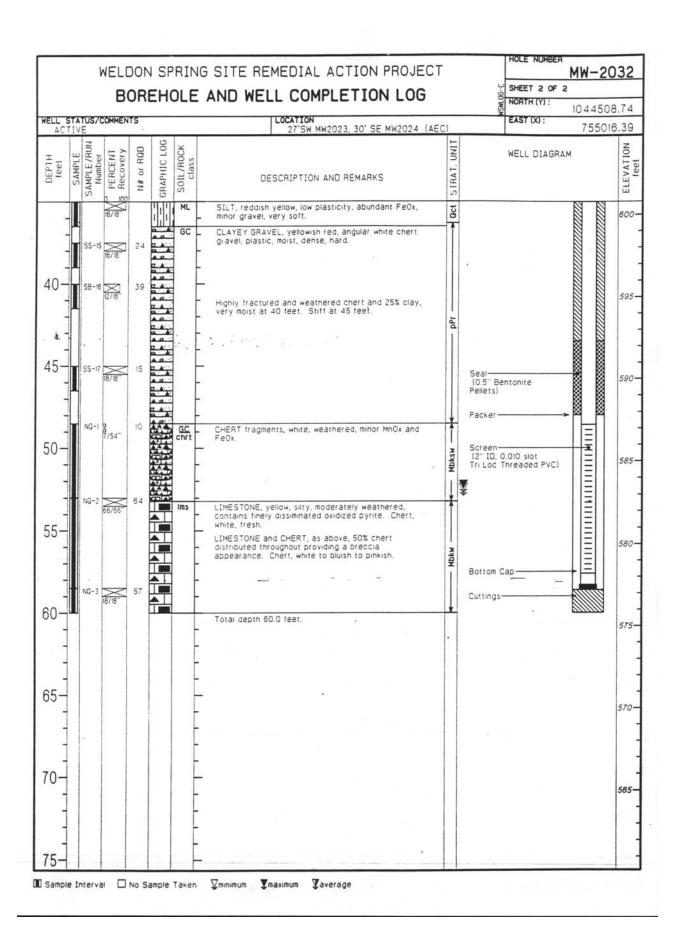
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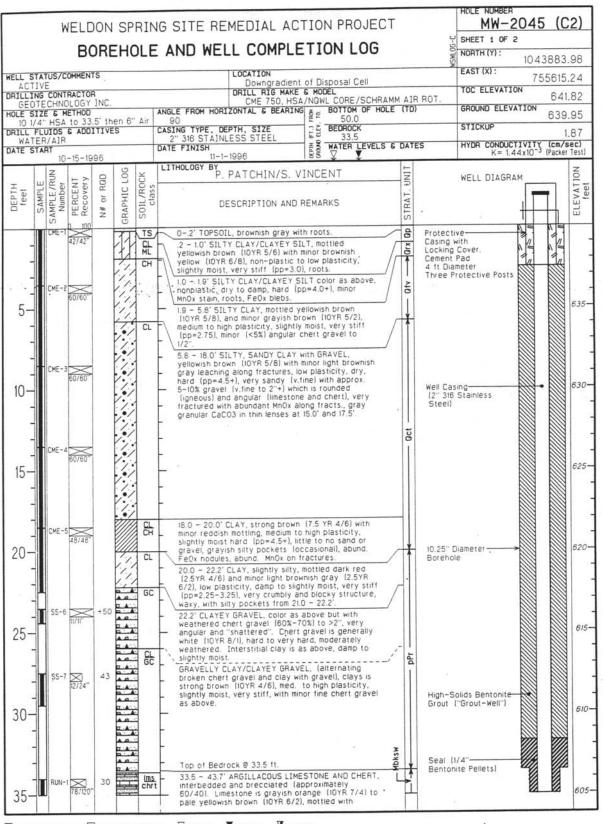
APPENDIX A

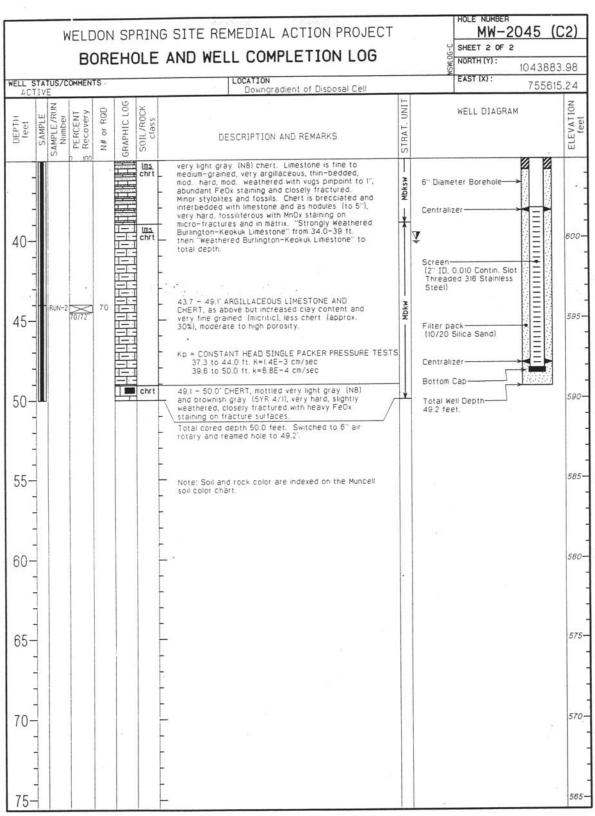
Monitoring Well Installation Logs

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_	4		10/10					1.00				
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	1	ST-6 SS-7	24/30	14			SILTY CLAY, brownish yellow, m					620
20-	1	ST-8	30/30"				gamp, hard. ft.		- 0ct			615
	ł	SS-9	18/18"	16		СН	SILTY CLAY, as above, highly p	lastic		Grout —		
25-	I	ST-10	11/12"			M.C.	SILT to SILTY CLAY, light brow plasticity, moist, very stiff. Min to 1.5 inches at 30 feet.	nish gray, medium or gravel, angular, up		(Volclay		610
	I	SS-II	18/18"	13		CH	CLAY, prownish yellow, highly pl stiff, contains 5% sand, 15% fine					
30- - -	I	ST-12	18/18"	34			Minor gravel, angular, up to 1.5 i	nches at 30 feet.				605
	1	SS-13	13/18"	18			-			191		
35-	Ч	SB-14	16/18"	33	-	ML	<u></u>		H		2	

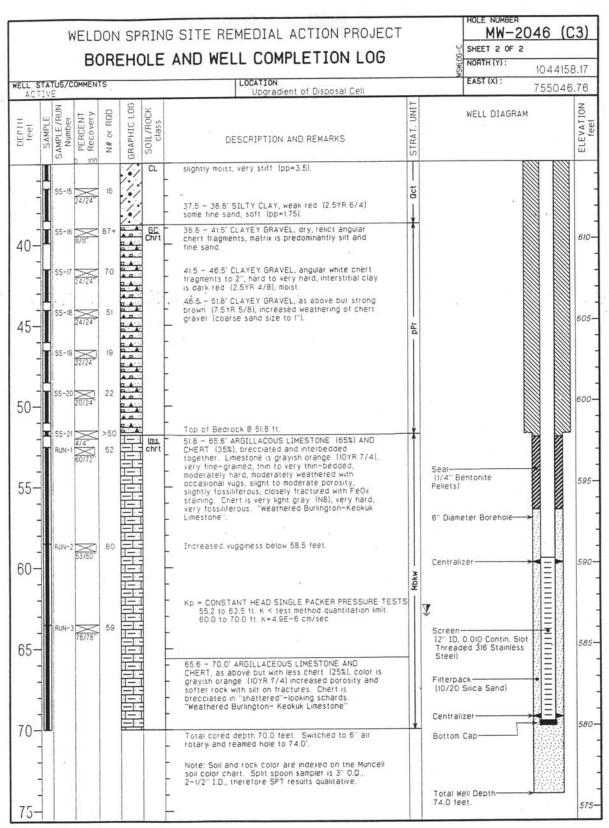
■ Sample Interval □ No Sample Taken ▼minimum ▼maximum ▼average



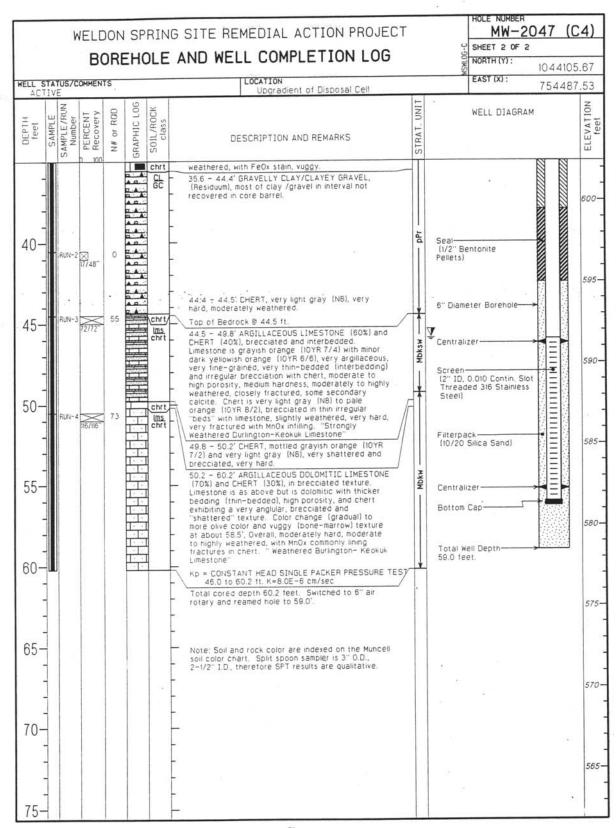




		SHEET 1 0		CTION PROJECT ETION LOG	WELL COMPL					W	
3.76	755046	EAST (X):		of Disposal Cell	LOCATION			S	MMENTS	us/c	L ST
.54	VATION 653	TOC ELEVA	A ATR	E & MODEL	DRILL RIG MAK		-	_	RACTOR	CONT	ACTI
7.77	ELEVATION 649	GROUND EL	(TD)	NG & BOTTOM OF HOLE	OM HORIZONTAL & BEARI	I A			LOGY	CHNC	GEOT
3.77		STICKUP		EBEDROCK	YPE, DEPTH, SIZE		en 6" /	the	to 51.8	HSA	10 1/
	NDUCTIVITY (cm/s	S HYDR COND	DATE	51.8 WATER LEVELS &	STAINLESS STEEL					/AIR	
	= 4.85x1U ° (Packel	K=	1. 1		11-15-1996	1		3	5-1996		E 31
10	AGRAM	WELL DIAG	UNIT	AUL PATCHIN	SEAN VINCENT/PA	5	507	ROD	- >	NO.	
ELEVATION	+		STRAT.	D REMARKS	DESCRIPTION AND	SOIL/ROCK class	GRAPHIC LOG	N# or R	PERCENT Recovery	SAMPLE/RUN Number	leel
		Protective————————————————————————————————————	V ⊞ X	(3" minus rock)	5' FILL, gravel drill pad	Fill	===		300		1
		Locking Cover. Cement Pad	*	ellowish brown (10YR	Y CLAY/CLAYEY SILT, ye	CI	7				7
	Posts	4 ft Diameter Three Protective Po	1	ist, very stiff	low plasticity, slightly mo 3.0).	불		19	18/24"	55-1	
				wish brown [10YR 5/4]	13.8: ELAY, mottled yellow grayish brown (10YR 5/2).	. []
645				stiff (pp=2.75).	icity, slightly moist, very			18	19/24"	55-2.	5_]
						1					J
								12	24/24"	55-3	
			50						24124		_
			11	taining is common.	, as above, black MnOx s			17		55-1	
640		Well Casing————————————————————————————————————		TO TO STATE OF THE				11	24/24"	55-4	n_
		Steel)									
				p=3.5).	, as above, very stiff (p)			20	24/24"	55-5	_
					4				24124		4
63			1	wish brown (10YR	- 32.0' SILTY CLAY, yello	다	14.	25	_	55-6	-
63				ticity, approx. 5%	with light greenish gray ures, medium to high plas	CH	12		24/24"		5-
3				(pp=4.5), grading to	lar chert gravel to I" with ded igneous gravel, hard	1	1/				+
				ent with depth	er silt and fine sand conti	1	1/	19	24/24"	55-7	4
						1	12				4
63		10000.00					1/	27	\sim	55-8	4
3		10.25" Diameter—— Borehole					1/2		24/24"		0-
							1/2	32	\leq	SC-0	+
							11	32	24/24"	22-8	-
							11				-
62			Gct	m plasticity, waxy	Y CLAY, as above, mediu earance, hard (pp=4.0).	CL	.67	32	24/24"	55-10	
				•		1	1		24/24		5-
						1	1%	27	_	SS-1	-
							1.		24/24"		-
	onite W	High-Solids Benton				1	1.				Ī
62		Grout ("Grout-well"				-	10	28	24/24	55-1	_
		4				1	1		100,70		0-
			-	shove but with red	240' 61 70 61 10	1	1%	28		SS-1	Ī
minim				es occasional ragments.	- 34.0' SILTY CLAY, as YR 5/6] color and include angular to angular chert f	1			24/24"		
6		0	Ш	ith light gray (10YR) - 37.5' SILTY CLAY, as wish brown (10YR 5/6) w silt in fractures, occasio	1		29	24/24"	SS-1	- 35-

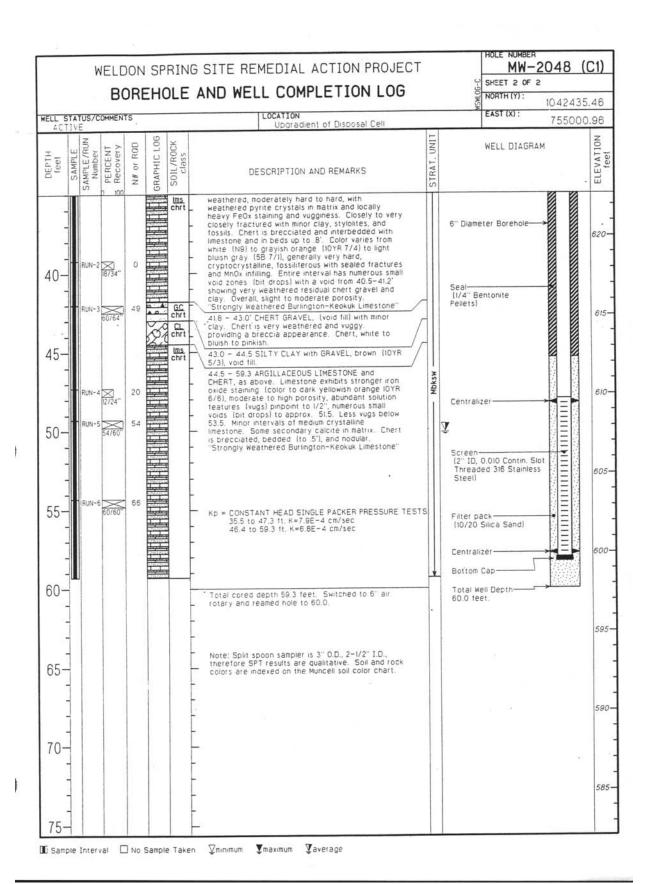


	_		_	_		Fire Post	1		HOLE NUMBER	7 (0	
		WELD	NOC	SP	RIN	G SITE REMEDIAL ACTION	PROJECT		MW-204	/ ((4)
		B	าคเ	= H() F	AND WELL COMPLETIO	NIOG	0-90	SHEET 1 OF 2 NORTH (Y):		
		ט	JIN	_110		AND WELL COM EE 120		MSM	10	44105	.67
		COMMEN	TS.			Upgradient of Disposi	al Cell		EAST (X): 75	54487	.53
DRILL		TRACTO				DRILL RIG MAKE & MODEL		ОТ	TOC ELEVATION	640	.30
		METHOD				CME 750, HSA/NGWL	OTTOM OF HOLE (TD)	01.	GROUND ELEVATION		
10	1/4" HS	A to 32	.0' th	en 6"			60.2 DROCK		STICKUP		
WA	TER/AI		21211			2" 316 STAINLESS STEEL	44.5		HYDR CONDUCTIVIT		2.82 sec)
DATE		2-3-19	96			12-10-1996	Ţ		HYDR CONDUCTIVIT K= 8.03x10	6 (Packe	
	Z	_ >	0	90	×	PAUL PATCHIN	LIND		WELL DIAGRAM		LEVATION
H +	SAMPLE/RUN	PERCENT Recovery	ROD	SRAPHIC LOG	SOIL/ROCK class						Peet .
DEPTH	SAMPLE/	ERC	io #N	APH	OIL/	DESCRIPTION AND REMARK	STRAT		Ī	-	ELEV
100000	SA	0 100	Z	6R	S					, i	ш
	П				Fill	0 - 4.5' FILL, 3/4" crushed rock to 1 ft, rock to 3.5', then silty clay fill with limest		rotective casing w			1
1	11			177	}	to 4.5'.		ocking (ement P	Cover.		1
	11				}		u 4	tt Diam			635-
	11.			-			- 111	illiee il	olectine rosis		
	55-	1 2	31	H	ML	SILT, light gray (2.5Y7/2), nonplastic, dr	, hard				1
5-		Lange Co.			,1000	(pp=4.5+) minor FeOx mottling.	, hard				1
	SS-	2 😾	24	17	ML	CLAYEY SILT, mottled yellowish brown (1 and light brownish gray (2.5Y 6/2), low p	asticity				-
		2 34/24"	276.0	[1]		slightly moist, hard (pp=4.5+), minor FeO	and ×				630-
	Щ			11		- MnOx.	1				1
	T SS-	3 16/24"	18	12	CL	SILTY CLAY, mottled light brownish gray and yellowish brown (10YR 5/6), low to m	(2.5Y 6/2)	dell Cook		. 📖	1 4
10-	1	10724		12		plasticity, slightly moist, very stiff to har	(pp=4.5), O		Stainless		4
	55-		22	1	CL	 minor silt as lenses in clay and minor fine gravel, Fe0x nodules and Mn0x streaking 		Steel)			4
	33	20/24"	22	1:/	"	SANDY CLAY with GRAVEL, mottled yellow	ish brown				625-
	1			./0/		(10YR 5/8) (primary) and minor light brow (10YR 6/2), slightly moist, sandy with app	msh gray				-
	SS-	5 🔀	16	(4)	1	gravel (very fine to 2"), mostly angular of	hert with	*			-
15-	-	20/24"		1%		minor subrounded igneous, abundant Fe0 stain, very stiff to hard (pp= 3.75-4.5+)	with				-
				1%		 near-vertical fractures exhibiting leaching color) and MnOx staining. 	g Igray				-
	-TI SS-	6 21/24"	22	1	CL	SILTY CLAY, distinctive yellowish-red (5	YR 4/6)				620-
	-			1/1	대	 moderate to high plasticity, slightly moist or sand, soapy, very hard (pp=4.5+) still 	no gravel				_
2	SS-		65	4	GC	(pp=3.0).	All				_
20-	-	23/24"		-		Harder drilling @ 19.0 ft. CLAYEY GRAVEL, approx. 75 - 80% ang		10.25" Di			
-	-14			4.0	-	gravel with plastic interstitial clay and m	nor sand	Doremone			
	_I SS	8 7/7"	>50	4.0		pockets. Gravel is exclusively chert until 28.0° then includes very weathered (to p	owder)				615-
	-				4	limestone. Interstitial clay is dusky red 4/4), high plasticity, slightly moist, with p	ockets of				
	- 55	.9	>50			distinctive yellow fine sand, abundant Fe minor MnOx stain on gravel, very dense.	Ox and				
25-	_ 4	11/11"		4.0	1						
1 -	4			4.0		-	pPr				
	-			4.0	3	Very nard drilling © 27 ft.					610-
1	50.	10	43	-	<u>.</u>	CLAYEY GRAVEL, as above, with weather	ed				,010
1	-	10 22/24"	1	7.4		limestone pieces and powder.			olids Bentonite		
30-				4.6		Sampler refusal @31.9 ft. Weathered lime	111	Grout ("Grout-Well")		
1 30				4.0	2	sampler. Auger refusal @ 32.0 ft. Starte					-
				-	ě	wireline coring.					3
	RUI	l−1 ⊠ 37/102	- 8		chrt	argillaceous limestone and as nodules to	.5' (minor)				605-
						chert is light gray (N7) to minor mottled gray (N5), slightly weathered, very hard	medium , very				'
35-						tossiliferous, closely fractured. Minor lim grayish- orange (10YR 7/4), argillaceou	estone is				'
1 22.						moderately hard, brecciated, moderately					



■ Sample Interval □ No Sample Taken ♀ minimum ♀ maximum ♀ average

						NG SITE REMEDIAL ACT		Т	,	SHEET 1 OF 2	2048	(C1
		B	DRE	-HC)LE	AND WELL COMPLE	TION LOG		CINCH	NORTH (Y):	104243	35.46
LL STA		COMMEN	TS.			LOCATION Upgradient of	Disposal Cell			EAST (X):	75500	0.96
ILLING	CON	TRACTO	TAIC			DRILL RIG MAKE	& MODEL	мм а1	R ROT.	TOC ELEVATION	in 65	59.85
LE SIZ	ZE &	METHOD			П	ANGLE FROM HORIZONTAL & BEARING	BOTTOM OF HOL	E (TD)	GROUND ELEVA	TION 6	57.72
10 1/4 ILL FL	UIDS	A to 32	3 th	en 6"	Air	CASING TYPE, DEPTH, SIZE	BEDROCK			STICKUP		2.13
WATE TE STA		R	- 2024 2 6		-	2" 316 STAINLESS STEEL DATE FINISH	32.3 ₩ATER LEVELS	& DAT	ES	HYDR CONDUCT	IVITY (cr	n/sec
		2-12-19	96			12-19-1996 LITHOLOGY BY	BS V V			K= 7.5	IIXIU - (Pac	ker Te
ш	NS.	<u>- ></u>	ROD	GRAPHIC LOG	č	PAUL PATCH	HIN	-INI		WELL DIAGRA	М	ELEVATION
feet	MPLE/RI	CEN	or RI	15	/RO ass						_	\ \ \
5 448	SAMPLE /RUN	PERCENT Recovery	* Z	RAPH	SOIL/ROCK class	DESCRIPTION AND F	REMARKS	STRAT.				ī
	S.	D 100	-	9				S	Barross	4		Ħ
					Fill	FILL, silty clay, dark brown, damp).	1	Protect Casing	with	3/1 3	/
				_		-		₹.	Locking Cement	Pad	147 14	7
	1							¥	4 ft Dia	meter rotective Posts	22 2	6
.]				12	. CL	SILTY CLAY, light brownish gray minor yellowish brown (10Y 5/6) n	(2.5Y 6/2) with nottling, medium	IT.				
_]	·SS-	15/24	9	11		plasticity, slightly moist, very stif	f (pp=3.25),					
5-				12		abundant reox noddes remmy on						
	55-	, 🗸	19	11111	СН	CLAY, as above but with increase	ed plasticity (high)	-11				
-	33	2 14/24"	,,,			and yellowish brown mottling, ver- decrease in silt content and no o	y stiff (pp=3.0),					1 6
-	Ų.											
-	55-	3 21/24	16			CLAY, as above, very stiff (pp+)	3.0).	710				
		21/24"				-		0	Well Ca:	sing————————————————————————————————————	111	
-	4					-			Steel)	O.C.		
-	SS-	4 22/24"	13			CLAY, as above, very stiff (pp=2)	2.75), minor coarse					
		23/24				_						6
	n ss-		12	11/1	CI	SILTY CLAY, as above, increase	in vellowish brown					
5-	22-	22/24"	12	1/1	다	color, MnOx streaking, and coarsi gravel content at bottom of sam	e sand and fine					
9				1/1		(pp=3.0).	pic, very surr					
	55-	6 23/24	19	6/	CL	SANDY CLAY with GRAVEL, yellow	wish brown (10YR	1				
		23/24"		17		 5/8) with minor light brownish gramottling, medium plasticity, slightl 	v moist, hard	- 11				1 6
	4			1.		(pp=4.5+), contains approx. 10% chert gravel (coarse sand size t	angular weathered to 1"), abundant					
	55-	7	18			MnOx streaks and FeOx stain.			10.25" [Diameter ——		
0-		C4154		(3)		SANDY CLAY with GRAVEL, as at moisture (moist), very stiff (pp=			Boreho			
-	-		15	1%		softer zones.		Det				
-	55-	8 24/24"	15	11111	Ci	CLAY, mottled strong brown 17.5	YR 5/6) and dark	$\dashv \bar{1}$				
-					다	red (2.5YR 4/6), medium to high sand, trace v. fine gravel, abund	plasticity, moist, no					
-	55-	9	18		CH	CLAY with GRAVEL, strong brown	(7.5YR 5/6) with					
5-		24/24"		1%	CH	minor light gray, medium to high p moist, very stiff (pp=3.5+), cont	plasticity, slightly ains no sand and					
-	4			1:/		 approx. 10-15% very fine to 1.5" igneousl, abundant MnOx streaks 	gravel (chert and	1				
-	SS-	10 19/24	30	0	CL	depth), last .5' very gravelly with	very abundant	1	1			
_		10/24		12%		MnOx and FeOx nodules. GRAVELLY CLAY, strong brown	(7.5 YR 5/6), high					
			+91	0		plasticity, approx. 40% angular minor limestone gravel, clay is mo	chert gravel and	1		olids Bentonite		
0-4	22.	9/14"	CANTO	63		nodules and MnOx streaks and co	onspicuous	PPr	Grout	("Grout-Well")		
٠ <u> </u>	SS-	12 11/18	24	4.0	CL GC	calcareous sand pocket (decom 26.5-27.0'.	hosed tock() at	/11				
	L	117.10		-	30	GRAVELLY CLAY, as above but s		/ []				
1	Pilit	-1	9		lms			/ 1	1			
1	NUN	36/78"	8		chrt		, hard, damp.	MDKsw				
_ 1						Top of Bedrock @ 32.3 ft.		/ Σ				
5-	-					32.3 - 41.8' ARGILLACOUS LIMES interpedded and preccieated.	Approximately 70%	1				
	_	1	1	_		ime-grained thin-bedded sligh en Tinimum Tinamum Tavers	e (10YP 7/4)	_	_			-



	W					IG SITE REMEDIAL AC		2	J-90	SHEET 1 OF 2	204	8 (a)
	STATUS/C				,,,,	LOCATION	DSAL CELL: PERIMETER	R WEL	MSM	NORTH (Y):			
DRILL	ING CONT	RACTOR TERN	R Inc.			DRILL RIG MAKE CME-750 HS/	ANOUL TO THE ON AT	R RO	TARY	GROUND ELEVA			
9"	SIZE & M HSA-34;	NQ-6	2: 6"		03	ANGLE FROM HORIZONTAL & BEARIN Vertical		on. W	ell	STICKUP	KITON		_
Wa	FLUIDS ter core:			S		CASING TYPE, DEPTH, SIZE 2" 316 SS Mon. Well DATE FINISH	E 2 32 E	DAT	ES		TIVITY	(cm/s	ec)
DATE	11-	9-01				12-5-01, Mon. Well	WATER LEVELS &			HYDR CONDUC K = 7.	91x10 ⁻⁴	(Packer	$\overline{}$
+	RUN	N 2	ROD	106	3CK	ALAN BENF	ER	UNIT		WELL DIAGRA	M	7	T10N
DEPTH	SAMPLE SAMPLE/RUN Number	PERCENT Recovery	N# or B	GRAPHIC LOG	SOIL/ROCK class	DESCRIPTION AND	REMARKS	STRAT.			#	H	ELEVATION feet
10-		3 100				Soil not sampled or logged from 28.0-1t.		Get/Grc undiff.	Well Cas 2" 316L Steel	orth Cover			-5
25-										"Grout-Well")			
30-	SPT-1		9		CH	CLAY, high plasticity, ~40% ang weathered limestone gravel, mo- moist, firm CH. Residuum. CLAY, gravelly to clayey gravel reddish brown, moist, hard, CH-	tly yellow brown, , high plasticity clay,	- pPr					-30-
35-	SPT-3	31/66"	76+		CHRI LMS	LIMESTONE, weathered, white t dry, hard. Auger refusal at 34.0-ft. Cont NO-1, 34.0° - 39.5°. Poor recove unknown. Lost circulation at 34	nue with NO core.	₩ ₩					

WELDON SPRING SITE REMEDIAL ACTION BOREHOLE AND WELL COMPLETIO	CHEET 2 OF 2
WELL STATUS/COMMENTS LOCATION SO. OF DISPOSAL CE	EAST (X):
SAMPLE ALL SAMPLE SAMPLE SAMPLE ALL SAMPLE A	METTO NIT
100 LIMESTONE, slightly weathered, crystalin locally argillaceous, with specks and thin locally argillaceous, but of "near bottom of core interbedded light gray chert. Rock in the feeth, hard, with minimal oxidation. Two onear core bottom; one is 3-1/2" thin ph plasticity, light plasticity	streaks of lowers in angular lower borehole 6" Diameter Borehole 6" Diameter Borehole 6" Diameter Borehole 6" Diameter Borehole 40— 40— 40— 40— 40— 40— 40— 40— 40— 40

		VEI D	100	SP	RIN	G SITE REMEDIAL ACT	TON PROJECT			HOLE NUMBER	MW	-20	51
	,								J-9	SHEET 1 OF			
		B	JKI	=HC	JLE	AND WELL COMPLE	TION LOG		OWSW	NORTH (Y):	104	2175.	.97
	STATUS/0	COMMEN	TS			LOCATION NORTHEAST EL	DGE OF DISPOSAL CEL	L		EAST (X):	7	53129	1.19
DRILL	TIVE					DRILL BIG MAKE 8			AIR ROT	TOC ELEVATI	ON	639.	.77
HOLE	YNE WES SIZE & M /4" Auge	METHOD		hon 6		ANGLE FROM HORIZONTAL & BEARING	BOTTOM OF HOLE	(TD)		GROUND ELEV	ATION	636.	.43
DRILL	FLUIDS	CODA 3	ITIV	ES	All	CASING TYPE, DEPTH, SIZE 2" 316L Stainless Steel	BEDROCK 26.5			STICKUP			34
DATE		-2-200	20			DATE FINISH 5-30-2000	WATER LEVELS & I	DATE	S	HYDR CONDUC	7x10 ⁻³	(cm/s (Packer	ec) Test)
	Z			90	¥	LITHOLOGY BY ALAN BENFER/PAUL		UNIT		WELL DIAGRA			Z C
H to	SAMPLE MPLE/RU Number	PERCENT Recovery	RGD	GRAPHIC LOG	SOIL/ROCK class	ALAN DEM ENVIOLE				WELL DIAGRA	```\ \	ī	EVATION feet
DEPTH feet	SAMPLE AMPLE/RI Number	ERC	N# or	APH	O1L/	DESCRIPTION AND R	EMARKS	STRAT.			#	Ц.	ELEV
	SA	0 100	Z	1		FILL. (silty clay)		E. S	Protecti		N=	N F	
	- 11			y/	CH	CLAY high plasticity mottled velic	ow brown and light	1	Casing	with	7	7/	635
	SPT-	2	5		CH	gray, approx. 10% fine white sand	i, moderate black		Cement 3 ft Dia	Pad	7/2	7/	-
	-111			1/		CLAY, high plasticity with some sa	and and gravel		Four Pro	otective Posts			-
-	SPT-	3	12	1/		, and black, moist, hard, GH.							
5-				1/		Lenses of clean fine sand betwee							
	SPT-	4	18	1/		CLAY, as above, approx. 10% fine fine gravel, hard, MnOx on vertical	sand and some I fracture						630-
	SPT-		12	1/		surfaces, CH.		Oct .	Well Cas		₩.		1
	SFI	0	12	1/				Ī	(2" 316) Steel)	L Stainless			
10-	SPT-	6	14	1/									-
	-			1/									625-
	SPT-	7	23	777	CL	SILTY CLAY, low plasticity, with w							-
15	-					nodules, light gray, moist, hard, CL	-		.0.05:: 5	Diameter			-
45	SPT-	8	14	11	СН	CLAY, high plasticity, FeOx nodule gravel, mottled light brown and lig	es, trace fine	¥	Borehol				
15-				14	СН	hard, CH. CLAY, as above but more reddish	'	1					
	SPT-	9	10	100		plasticity, with some angular cher	t gravel, CH.						620-
	- SPT-	10	В	1	1			Ш					
91				69	1			Ш	High-S	olids Bentonit e "Grout-Well")			
20-	SPT-	n	7	89	1	-		1	0.000	0.001			-
1.00	-			9	1			- pPr					645-
	SPT-	12	15	K									-
			l'i	6				Ш					-
25-	SPT-	13	19	169	1		1(*)	Ш					1 1
23	_11					- Sampler refusal at 25.9 ft. Auger	red to top of rock						610-
1	- RUN-	38/60"	8		chrt	at 26.5 ft (auger refusal). 26.5' - 35.3' CHERT AND ARGILLA	ACEOUS	1					-
	-	50700			1	LIMESTONE (approx 65%/35%), C gray (NB) to minor medium light g	thert is very light gray (N6),	Ш					_
	-					 brecciated to very thinnly bedde very hard, slightly to moderately 	d with limestone, weathered,						-
30-					i	 fossiliferous, with MnOx and FeOx microfractures in chert). Limesto 	ne is grayish	MDKW	Can'				-
	RUN-	2	46	1		- orange (10YR7/4), mod. weather with vugs pinpoint to 3 mm, fine g	rained, very thinly	Σ		Enviroplug			605-
		2 47/60				 bedded (interbedded with chert) fossiliterous with FeOx staining (in possible void zones at 28.0° – 28 	particularly in		benton	ire minst			-
						34.7').	310 010						
35	_Ш			-	4_	Lost circulation at 29.0° Possible voids from 34.0 - 34.7°		Щ					
1						1 035,010 10103 11011 3410 3417				100			1

WELDON SPRING	SITE REMEDIAL ACTION PROJECT		TOTAL CONTRACTOR OF THE PARTY O	r-2051
BOREHOLE A	AND WELL COMPLETION LOG	9	SHEET 2 OF 2 NORTH (Y):	42175.97
WELL STATUS/COMMENTS	NORTHEAST EDGE OF DISPOSAL CELL		FAST (X) ·	53129.19
DEPTH feet SAMPLE FRUN Number PERCENT PERCOVETY N# or RQD GRAPHIC LOG SOIL/ROCK class		STRAT, UNIT	WELL DIAGRAM	ELEVATION
10 AND 10	35.3' - 43.7' ARGILLACEOUS LIMESTONE (70%) and CHERT (30%) as above but increased limestone. Inreased vugs (to i") parallel with bedding, incoderately weathered. Probable voids @34.0' - 34.7' and 36.5' - 38.0' with neavy Fe0x staining and residual chert gravel. Semi-fresh pyrite specks @33.5'. Mnox specks @33.5'. with nereased Fe0x and weathering. Yellowish coloring (clay) at 40.3' - 41.5' with nematite blebs.	Screen—2" (10 Si Continuos	iot) 316L SS us Wrap killica Sand)	595— 595— 585— 580— 575—
70-				565-

Ⅲ Sample Interval □ No Sample Taken 및 minimum 및 maximum 및 average

						_				HOLE NUMBER	CALL STATE	1525(200)	27/27
		W	FID	ON	SP	RIN	G SITE REMEDIAL ACTION PROJECT	ŗ.			MW-	-205	55
		**							3-9	SHEET 1 OF 2			
			BC	JKE	HC	ILE	AND WELL COMPLETION LOG		O IMS	NORTH (Y):	104	2419.	92
WELL S	STAT	'US/C	OMMENT	S			LOCATION CELL PERIMETE	D ME	3	EAST (X):	755	000.	58
	TIVE	=					SO. OF DISPOSAL CELL; PERIMETE DRILL RIG MAKE & MODEL			TOC ELEVATION	17.00	662.	
LAY	YNE	WES!	TERN]	Inc.		_	CME-750 HSA/NOWL; 1-R TH-60 A	IR RI	D)	GROUND ELEVA	TION		
HOLE S	HSA	-34	NQ-62	2: 6"	AJR-	63 1	Vertical			STICKUP		659.	
DRILL	FLU ter (DIDS ore:	CODA 3	ITIVE am	S		01 316 CC Nee Well 22 5	0.04	****		TVTTY		76
DATE S		RT.	9-01				DATE FINISH 12-5-01, Mon Well WATER LEVELS	& UA	IES	HYDR CONDUCT	5x10 ⁻⁵	(Packer	Test)
	П			0	90	¥	LITHOLOGY BY ALAN BENFER	- LINI		WELL DIAGRA	мг	_	No.
Ξ_	SAMPLE	/RU	PERCENT Recovery	ROD	GRAPHIC LOG	SOIL/ROCK class		75		MELL DIAGNA		111	EVATION feet
DEPTH	AMP	MPLE/R Number	RCE	0	PHI	1L/F	DESCRIPTION AND REMARKS	STRAT			#	H	E E
"	S	SAM		* Z	GRA	SO	MALE AND IN THE COLUMN TO THE	ST			1		EL
-	+	1200	100					1	Protect		3		-
	11							- 11	Locking	Cover.			-
	11						Soil not sampled or logged from the surface to 28.0-ft.	Ш		te Pad With			4
	11				0				4 Prote	ctive Posts			107.2
	11						- 110 M HAVE X 120 F						655-
5-	-												_
	-11						-						
	41						-						
	-11						- 5						
	-11						-		Well Ca 2" 316L	sing————————————————————————————————————	1		650-
10-	41						_		Steel				650-
"	41						-						1
	1						_	1	2				-
	1							a point					-
]].							3					-
15-	1							040/400					645-
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	1								Boreho	le			-
1 2													
100	7												640-
20-													
	1												
	1												
1	1												
	1								High-9	Solias Bentonite			635-
25-	-						<u></u>			("Grout-Well")			
1	-												`.
	-						†			-			
	-	SPT-	1	9	V	CH CH	CLAY, high plasticity, ~40% angular chert and	T	Ħ				
	-11				1/2	1	 weathered limestone gravel, mostly yellow brown, moist, firm CH. Residuum. 						630
30	-#	SPT-	2	14	R	1	CLAY, gravelly to clayey gravel, high plasticity clay.		pPr				630-
	-	10000000			1		reddish brown, moist, hard, CH-GC. Residuum.		1				
1	-	SPT-	3	76+	1	2	-		¥				
	-	4		100	H	CHR			*				
	1	NG-	D	15	H.		dry, hard. Auger refusal at 34.0-ft. Continue with NQ core.	H	¥QK				7
35			31/66"	,,,	缸	4_	NO-1 340'- 39.5'. Poor recovery, loss zones	-	Ц		8	- 12	625-
1 33							unknown. Lost circulation at 34.5' permaently.						

	V	VELD	100	N SP	RIN	SITE REMEDIAL ACTION PROJECT			HOLE NUMBER	MW-20	055
		В	OR	EHO	DLE	AND WELL COMPLETION LOG		J-907-M	SHEET 2 OF 1	104241	10.02
	TATUS/C	OMMEN	TS			SO. OF DISPOSAL CELL; PERIMETER	WF	<u>§</u>	EAST (X):	75500	
DEPTH Feet 57	SAMPLE /RUN MANUMBER	PERCENT Recovery	N# or RGD	GRAPHIC LOG	SOIL/ROCK class	DESCRIPTION AND REMARKS	STRAT. UNIT		WELL DIAGRA	м	ELEVATION
40-	NO-2	28/40" 52/70" 52/70"	62 !3		CHRT	LIMESTONE, slightly weathered, crystalline-grained, locally argillaceous, with specks and thin streaks of MnOx, fossinferous, light gray to light yellowish brown, vugs up to 1" near bottom of core, some interbeded light gray chert. Rock is relatively fresh, hard, with minimal oxidation. Two clay seams near core bottom; one is 3-1/2" thick with angular chert fragments, lower one is 6" of massive clay with some tine gravel, high plasticity, yellow brown, minor MnOx. Burlington-Keokuk Limestone. 8 39.5'- 42.8'. LIMESTONE as above, becoming oxidized at -39.7', some stypolites, -30% scattered chert, appears brecciated, fossiliterous, mostly light gray. Up to 4+ fractures per foot, rough, open, oxidized. 8 '41.5'. Scattered oxidized sulfides, increased limestone weathering at 42.7'. Bit dropped at ~41.8' to ~42.5'. Weathered Burlington-Keokuk Limestone. 8 '43.5'- 44.1'. CHERT, hard, white, light and blue gray. 8 '44.6'- 45.0'. Silty clay with limestone fragments. 8 '45.4'- 48.6'. LIMESTONE, moderately weathered, moderate solutioning with vugs up to 1", generally coarse-grained, some argillaceous, FeOx staining with thin streaks of MnOx, mostly light yellow brown, with ~ 30% scattered interbedded chert, generally light gray. 8 '48.2'- 52.0'. CHERT, hard, with thin streaks of MnOx, fracture surfaces oxidized, white to pale yellow, with up to ~25% thinly interbedded immestone, strongly weathered at 49.2-ft., soft, argillaceous, dark yellowish orange. At 50.1 to 52.0-ft., limestone is very vuggy along bedding: clayey. One to 4+ fractures per foot, rough; broken. Strongly weathered Burlington-Keokuk Limestone. 9 52.0'- 53.8'. CHERT, hard, bluish gray with ~40% limestone, strongly weathered, argillaceous, generally interbedded, a few vugs, yellowish orange. 9 ~53.7'. Cut fast. Very vuggy at 53.8-ft. CHERT AND LIMESTONE as above. ~40% limestone, strongly weathered, locally can be scratched with fingernal, a few vugs. Chert is bluish gray, locally fossiliferous. One to 3+ fractures per foot, rough, open, some broken	MDK5W — MDK MK MK	Centrali Centrali Centrali Centrali Screen Continu Filterpa 10/20 S Central Bottom Total W 62.0-ft	viroplug te Chips zer tic Water Level t. Slot) 316L SS bus Wrap ck ck ck Cap And ell Depth		610-
65-						6" diameter to 63.0" and a 2" monitoring well was constructed.		6" Hole 63.0-ft			595•
70-						Packer testing was not performed.					590

 ${
m I\!C}$ Sample Interval \square No Sample Taken $\sqrt[n]{2}$ minimum $\sqrt[n]{2}$ maximum $\sqrt[n]{2}$ average

APPENDIX B

Statistical Evaluation of Detection Monitoring Data (1998 to 2001)

Statistical Evaluation of Detection Monitoring Data

B.1 Evaluation Summary – 1998 through 2001

Under the original version of this plan, the elements of the detection monitoring program included:

- Collecting four replicate samples at each location on a semiannual basis,
- Measuring groundwater elevation at each well location, as well as flow rate for the spring, on a quarterly schedule and immediately prior to each semiannual sampling event,
- Analyzing for the entire list of constituents presented in Table 3-2 of the main text of this report, and noting any unusual colors, odors, or turbidity,
- Evaluating analytical data in comparison with background levels to identify statistically significant increases that may indicate an impact from the disposal cell, and
- For parameters that appear to exceed background levels: reviewing analytical results for potential errors, evaluating cell leachate volume data to confirm liner integrity, and resampling individual locations for the suspect parameters.

The detection monitoring data obtained from 1998 to 2001 were evaluated in accordance with the U. S. EPA guidance on *Statistical Analysis of Groundwater Monitoring Data at RCRA Facilities* (Ref. B-1). This document provides guidance on conducting various types of statistical analyses under the RCRA groundwater monitoring regulations (40 CFR 264, Subpart F). The foundational assumption of each statistical method is that the waste management unit is situated on a uncontaminated aquifer and that the only source of increases in contaminant concentrations in the groundwater is leakage from the waste management unit. The guidance cautions against the use of the prescribed methods in evaluating data from wells that have shown evidence of preexisting contamination or where a high degree of spatial variation exists between the background wells and compliance wells, both of which are true for the Weldon Spring site.

In the absence of regulatory guidance on more appropriate statistical methods for use at a site with preexisting groundwater contamination, detection monitoring data have been evaluated by several different methods, as discussed below.

B.1.1 1998 Results

Detection monitoring data from 1998 were evaluated by means of both parametric and nonparametric analysis of variances (ANOVA) analyses. Results of these analyses, which are

presented in the *Weldon Spring Site Environmental Report for Calendar Year 1998* (Ref. B-2), are based on a comparison of data from the compliance wells, MW-2032 and MW-2045 through MW-2047, with data from the upgradient (i.e., "background") well, MW-2048. These analyses resulted in a large number of statistical failures which, if they had been based on data from a previously uncontaminated aquifer, would have provided evidence of groundwater impact due to the disposal cell.

Many of the test failures were determined to be attributable to preexisting concentrations of certain parameters being higher in the compliance wells than in the upgradient well prior to waste placement (March 1998). However, after disregarding the parameters in which this was the case, the following parameters still failed at least one of the statistical tests:

- MW-2032 Chromium, silver, thallium
- MW-2045 Calcium, radium-228
- MW-2046 Silver, vanadium, TOX
- MW-2047 Vanadium, zinc, 1.3.5-TNB

The monitoring data for parameters that failed the interwell comparisons were further evaluated by means of ANOVA procedures based on intrawell comparisons with baseline data from the same locations. This testing resulted in the following statistical failures:

- MW-2045 Calcium
- MW-2046 Vanadium
- MW-2047 Vanadium, 1,3,5-TNB

All of the above statistical failures were attributed to natural fluctuations in the existing groundwater quality. It was not reasonable to consider these test failures to be indicators of cell leakage because waste placement, and subsequent leachate production, began only a few months before the first 1998 detection monitoring event, and contaminant fate and transport analyses had predicted a 53-year interval before contaminants leaking from the cell would be detected in the monitoring wells (Ref. B-3). In addition, the use of the upgradient well, MW-2048, as a "background" well was determined to be inappropriate since several constituents were already higher in this well than in any of the compliance wells before waste placement began.

B.1.2 1999 Results

The detection monitoring program was modified in 1999, after review of the previous two years of groundwater and leachate data. Several parameters were eliminated from the monitoring list. Also, the monitoring frequency was reduced to a single sample obtained semiannually from each location instead of the four replicates previously collected.

In an effort to derive a more reliable means of evaluating data, an intrawell tolerance interval approach was used to evaluate the 1999 data instead of the ANOVA procedures used the

previous year. A intrawell tolerance limit approach was considered the preferred method of evaluating data because this approach resulted in fewer false positive results that any of the other types of statistical analyses performed to date. Also, due to the heterogeneous nature of the aquifer it can be expected that each well would act independently because it monitors a discrete portion of the aquifer. By this method, each monitoring location (including the upgradient well) was considered to be a point of compliance, and "background" conditions were described by the contaminant concentrations measured at each location during baseline monitoring. Tolerance limits were calculated for each parameter at each monitoring location according to the methodology in *Statistical Analysis of Groundwater Monitoring Data at RCRA Facilities* (Ref. 10).

Using the baseline data collected prior to waste placement, upper tolerance limits were established based on the assumptions of a normal data distribution and a 95% level of confidence. Data from the two semiannual monitoring events were compared to the baseline values, and any exceedances were investigated through the data verification process, sample reanalysis, and/or resampling. All confirmed exceedances were reported as statistically significant increases. The *Weldon Spring Site Environmental Report for Calendar Year 1999* (Ref. B-4) summarizes the results of statistical analysis of the 1999 data, in which the following parameters exceeded baseline for at least one of the sampling events:

- MW-2032 Chemical oxygen demand (COD), chromium, nickel
- MW-2045 Arsenic, chromium, molybdenum, nickel
- MW-2046 Aluminum, barium, chromium, magnesium, nickel, COD
- MW-2047 COD
- MW-2048 Magnesium, sulfate

As in the previous two years, the above statistical failures were attributed to natural fluctuations in the existing groundwater quality. However, in accordance with the original version of this plan, a thorough study of the monitoring network was conducted in 2000 to confirm that the recurring baseline exceedances were not true indicators of cell leakage. This study was documented in the *Weldon Spring Site Cell Groundwater Monitoring Demonstration Report* (Ref. B-5). It included an evaluation of historical site-wide groundwater quality, review of leachate flow rate and analytical data, analysis of groundwater elevation fluctuations, comparison of filtered and unfiltered samples for metals analysis, and review of cell well construction and performance information.

The demonstration report concluded that the baseline exceedances were not due to contaminant migration from the cell, but rather were the result of variations in previously existing groundwater contamination compounded by poor hydraulic performance of some of the wells. The following actions were recommended to alleviate the recurrence of similarly false positive results in future sampling events:

- Attempt to improve the flow rate and clarity of groundwater in MW-2045 by redeveloping it prior to the next sampling event,
- Install an additional compliance well in the vicinity of MW-2045 to provide supplemental monitoring on the northeast side of the disposal cell, and
- Recalculate the upper tolerance limit for the baseline values of each parameter at each well. The new limits should be based on the assumption that the four replicates obtained during each quarterly baseline event were not truly independent samples but represented a single event.

Results of the filtered metals analyses confirmed that most of the metals exceedances coincided with high turbidity and likely resulted from metals adhering to suspended clay particles in the groundwater. Although the filtering of groundwater samples for metals analyses is an acceptable sampling procedure, it was not listed as a recommendation in the demonstration report because baseline values were already established using unfiltered samples.

B.1.3 2000 Results

The recommendations from the demonstration report were implemented, and the 2000 data were evaluated using the tolerance interval approach with the recalculated tolerance limits. The *Weldon Spring Site Environmental Report for Calendar Year 2000* (Ref. 13) contains the results of this evaluation, in which the following parameters exceeded the new baseline tolerance limits during at least one of the semiannual sampling events:

•	MW-2045	Chromium,	molybdenum

- MW-2046 Molybdenum
- MW-2047 Chromium
- MW-2048 Chromium, magnesium, molybdenum, sulfate

B.1.4 2001 Results

Results of the 2001 detection sampling, which were evaluated in the same manner as in the previous year, are presented in the *Weldon Spring Site Environmental Report for Calendar Year 2001* (Ref. 14). The following parameters were identified as exceeding baseline tolerance limits during at least one of the semiannual sampling events:

- MW-2045 Chromium, molybdenum, nickel
- MW-2046 Nickel, 2,4,6-TNT
- MW-2048 Sulfate

Two new wells were installed and one was abandoned under the disposal cell monitoring program in 2001. MW-2051 and MW-2055 were installed and MW-2048 was abandoned.

Baseline monitoring data was collected from these wells in 2001 and 2002, and they were added to the detection monitoring program in 2002.

B.2 Evaluation Summary –2004

In response to a comment from the MDNR regarding the distribution of the groundwater data from the disposal cell wells, a statistical evaluation of the data was performed. This analysis consisted of a determination of the data distribution and the appropriateness of the baseline tolerance limits for evaluation of the detection monitoring data.

B.2.1 Data Distribution

The data for the signature parameters at locations MW-2032, MW-2046, MW-2047, MW-2051, MW-2055, and SP-6301 were reexamined to determine whether the data is Normal or log-Normal. Testing for Normality or log-Normality were done by three (3) different methods, as suggested as alternative tests in the *EPA Statistical Analysis of Ground-Water Monitoring Data at RCRA Facilities – Addendum to Interim Final Guidance* (Ref. B-6). These tests were:

- 1. Probability Plot Correlation Coefficient
- 2. Shaprio-Wilk Test of Normality (n<50) or Shaprio-Francia Test of Normality (n>50)
- 3. Coefficient of Skewness

The tests were performed for both the non-transformed data and log-transformed data for each of the four signature parameters at each location. Each of the signature parameters at each of the locations passed at least one of the three tests for Normality and log-Normality. For example, at location MW-2051 the results were:

Analyte			Test M	lethod		
7 thatyte	PPCC	CS	SW/SF	PPCC	CS	SW/SF
		n-transformed [Data	Log	g-transformed D	ata
Barium	N		N	N	N	N
Iron	,-	N			N	
Manganese	N	N	N	N	N	N
Uranium	N		N	N	N	N

PPCC - Probability Plot Correlation Coefficient

CS - Coefficient of Skewness

SW - Shapiro-Wilk Test of Normality (n<50)

SF - Shapiro-Francia Test of Normality (n>50)

N - Criteria for Normality met

The other locations show similar results. Although the data shows a slightly stronger evidence of log-Normality than Normality, the data can be treated as Normal because of the difficulty in calculating the mean and variance/standard deviation for a log-Normal distribution.

B.2.2 Review of Baseline Tolerance Limits

All the available data was used in calculating baseline tolerance limits. Data points that may have been compromised in some manner should be excluded. Compromised data may include data collected after any disturbance of the sub-surface such as by drilling, excavation, soil sampling, etc. that may dramatically increase the mobility/solubility of some contaminants.

To demonstrate that there is little difference in the method used to calculate the baseline tolerance limit, values for the signature parameters at three of these locations were calculated using six methods (Table B-1). All of the data for each location were used in the calculations.

Table B-1	Calculated Baseline	Tolerance Limits for	or MW-2032,	MW-2046,	and MW-2051
lable B-1	Calculated baseline	Tolerance Limits it	JI 1010 V-2032,	IVIVV-ZU-O,	and wive Z

Location	Method	Ba (μg/I)	Fe (μg/I)	Mn (μg/l)	U (pCi/l)
MW-2032	EPA Guidance - Normal Data (a)	338.8	889.9	45.2	5.60
	Tolerance Limit – Normal Data (b)	376.7	1125.2	56.6	6.42
	xbar+3s - Normal Data	389.9	117.8	56.3	6.73
	EPA Guidance – log-Normal Data (a,c)	334.2	926.7	45.7	6.96
	Tolerance Limit – log-Normal Data (b,c)	370.7	1178.1	57.4	9.35
	xbar+3s - log-Normal Data (c)	383.4	1170.3	57.0	9.89
	EPA Guidance - Normal Data (a)	256.5	1238.6	147.7	1.67
	Tolerance Limit - Normal Data (b)	276.7	1577.5	186.9	1.76
	xbar+3s - Normal Data	287.0	1566.9	185.7	1.84
MW-2046	EPA Guidance – log-Normal Data (a,c)	249.9	1156.0	151.2	1.48
	Tolerance Limit – log-Normal Data (b,c)	268.3	1464.2	191.9	1.92
	xbar+3s - log-Normal Data (c)	277.6	1454.6	190.7	2.02
MW-2051	EPA Guidance - Normal Data (a)	253.2	2200.8	205.5	3.68
	Tolerance Limit - Normal Data (b)	285.3	2895.9	265.4	4.51
	xbar+3s - Normal Data	236.4	1657.9	158.7	3.12
	EPA Guidance – log-Normal Data (a,c)	248.5	1384.8	286.7	3.27
	Tolerance Limit - log-Normal Data (b,c)	278.9	1799.4	374.9	4.64
	xbar+3s - log-Normal Data (c)	232.6	1061.0	217.8	3.20

a Calculated by method outlined in EPA Addendum to Interim Final Guidance for Statistical Analysis of Ground-Water Monitoring at RCRA Facilities

The method outlined in the EPA Guidance (Refs. B-1 and B-6) is designed to treat below detection limit values differently from other methods of calculating a benchmark or baseline tolerance limit where below detection limit values are typically set at one-half the detection limit. However, the EPA Guidance method assumes that all the below detection limit values have the same detection limit, which is seldom the case and complicates the analysis.

The values in the table for MW-2051 show more variation than the other locations, particularly for iron, manganese, and uranium. This is likely due to the small data sets, where only 5 or 6 values for each of the signature parameters have been collected, and one or two extreme or outlier values can skew the calculated value.

b Bowker, Albert H. and Gerald J. Liberman, Engineering Statistics, Section 8.12 and 8.13.

c Mean and standard deviation for log-Normal calculated by method from Gilbert, Richard O., Statistical Methods for Environmental Pollution Monitoring, Section 13.1.1.

Comparison of the six different calculation methods yielded the following conclusions:

- 1. There is not much difference in the EPA Guidance Normal Data values and the EPA Guidance log-Normal Data values except for iron and manganese at MW-2051. Although the below detection limit values are treated the same in both these calculations, the difference is likely due to small sample size and outlier values as noted above.
- 2. The EPA Guidance Normal Data values and the Tolerance Limit Normal Data show some variation in many cases. The difference is probably attributable to the difference in the treatment of below detection limit values. The same argument can be stated for the EPA Guidance log-Normal Data values and the Tolerance Limit log-Normal Data values.
- 3. The Tolerance Limit Normal Data values and the "xbar+3s" Normal Data values are very similar, except at MW-2051. This is expected because the only difference is the tolerance factor multiplier. The Tolerance Limit is calculated as "xbar+ks", where the tolerance factor multiplier 'k' is from a table depending on the sample size and the probability that the calculated interval contains a give percent of the distribution. For the "xbar+3s" method the multiplier factor is always 3. The range for this factor is from approximately 2.2 to 10.5. As the sample size decrease the tolerance factor multiplier increases. This accounts for the difference in the values at MW-2051. The same argument can be stated for the Tolerance Limit log-Normal Data values and the "xbar+3s" log-Normal values.

Based on the analysis discussed above, it was not recommend to change the method currently used (tolerances limits) for calculation of benchmarks for the signature parameters. All of the available data that has not been compromised should be used. In addition, the 'arithmetic mean plus 3 standard deviations' is appropriate for the non-signature parameters since they are not a concern in the leachate.

B.5 References

- B-1 U.S. Environmental Protection Agency, Office of Solid Waste, Waste Management Division. *Statistical Analysis of Ground-Water Monitoring Data at RCRA Facilities Interim Final Guidance*. EPA/530-SW-89-026. Washington, D.C. April 1989.
- B-2 MK-Ferguson Company and Jacobs Engineering Group. *Weldon Spring Site Environmental Report for Calendar Year 1998.* Rev. 0. DOE/OR/21548-773. Prepared for the U.S. Department of Energy, Oak Ridge Operations Office. Weldon Spring, Missouri. July 1999..
- B-3 Tomasko, D., J. J. Quinn, L. A. Durham. *Groundwater Flow at the Chemical Plant Area of the Weldon Spring Site: Simulations of the Effect of the Disposal Cell on the Flow Field.* Prepared by Argonne National Laboratory, Environmental Assessment Division, Argonne, IL. October 1996.
- B-4 MK-Ferguson Company and Jacobs Engineering Group. *Weldon Spring Site Environmental Report for Calendar Year 1999*. Rev. 0. DOE/OR/21548-845. Prepared for the U.S. Department of Energy, Oak Ridge Operations Office. Weldon Spring, Missouri. July 2000.
- B-5 MK-Ferguson Company and Jacobs Engineering Group. *Weldon Spring Site Cell Groundwater Monitoring Demonstration Report*. Rev. 0. DOE/OR/21548-864. Prepared for the U.S. Department of Energy, Oak Ridge Operations Office. St. Charles, MO. November 2000.
- B-6 U.S. Environmental Protection Agency, Office of Solid Waste, Waste Management Division. *Statistical Analysis of Ground-Water Monitoring Data at RCRA Facilities Interim Final Guidance*. EPA/530-SW-93-001. Washington, D.C. November 1992.

TABLE 5. TOLERANCE FACTORS (K) FOR ONE-SIDED NORMAL TOLERANCE INTERVALS WITH PROBABILITY LEVEL (CONFIDENCE FACTOR)

Y = 0.95 AND COVERAGE P = 95%

n	K	111	n	K
5 6 7 8 9 10 11 12 13 14 15 16 17 18	7.655 5.145 4.202 3.707 3.399 3.188 3.031 2.911 2.815 2.736 2.566 2.523 2.486 2.543 2.423 2.396 2.371 2.350 2.329 2.329 2.329 2.220 2.166 2.092 2.092 2.065 2.017 2.000 1.986		75 100 125 150 150 150 225 250 250 350 350 350 350 350 350 350 350 350 3	1.972 1.924 1.891 1.868 1.850 1.836 1.824 1.806 1.799 1.792 1.777 1.773 1.769 1.763 1.760 1.757 1.750 1.750 1.754 1.750 1.748 1.746 1.744 1.742 1.740 1.739 1.737 0.736 1.731 1.732 1.731 1.729 1.728 1.727

<u>SOURCE</u>: (a) for sample sizes \leq 50: Lieberman, Gerald F. 1958. "Tables for One-sided Statistical Tolerance Limits." *Industrial Quality Control*. Vol. XIV, No. 10. (b) for sample sizes \geq 50: K values were calculated from large sample approximation.

U.S. Environmental Protection Agency, Office of Solid Waste, Waste Management Division. *Statistical Analysis of Ground-Water Monitoring Data at RCRA Facilities - Interim Final Guidance*. EPA/530-SW-89-026. Washington, D.C. April 1989.

APPENDIX C

Groundwater Flow Rate Determination

Groundwater Flow Rate Determination

Groundwater flow rates and flow directions will be evaluated annually as specified in Section 4.4 of the main text of this report. Results for 1998 through 2002 are presented in this appendix.

C.1 Groundwater Flow Direction

The groundwater flow direction was determined by constructing a potentiometric surface map of the shallow aquifer using the available wells at the chemical plant (Figure C-1). Potentiometric surface maps (Figures C-2 through C-6) were constructed using the average of the groundwater elevations measured during each year. A summary of the average groundwater elevations for each well is included in this appendix.

The potentiometric surface has remained relatively unchanged from 1998 through 2002. The groundwater flow direction is to the north. A groundwater divide is present along the southern boundary of the chemical plant site.

C.2 Groundwater Flow Rates

The calculation of the average groundwater flow rate (average linear velocity) is a function of the hydraulic conductivity (K), the hydraulic gradient (I) and the effective porosity (n_e) of the shallow aquifer:

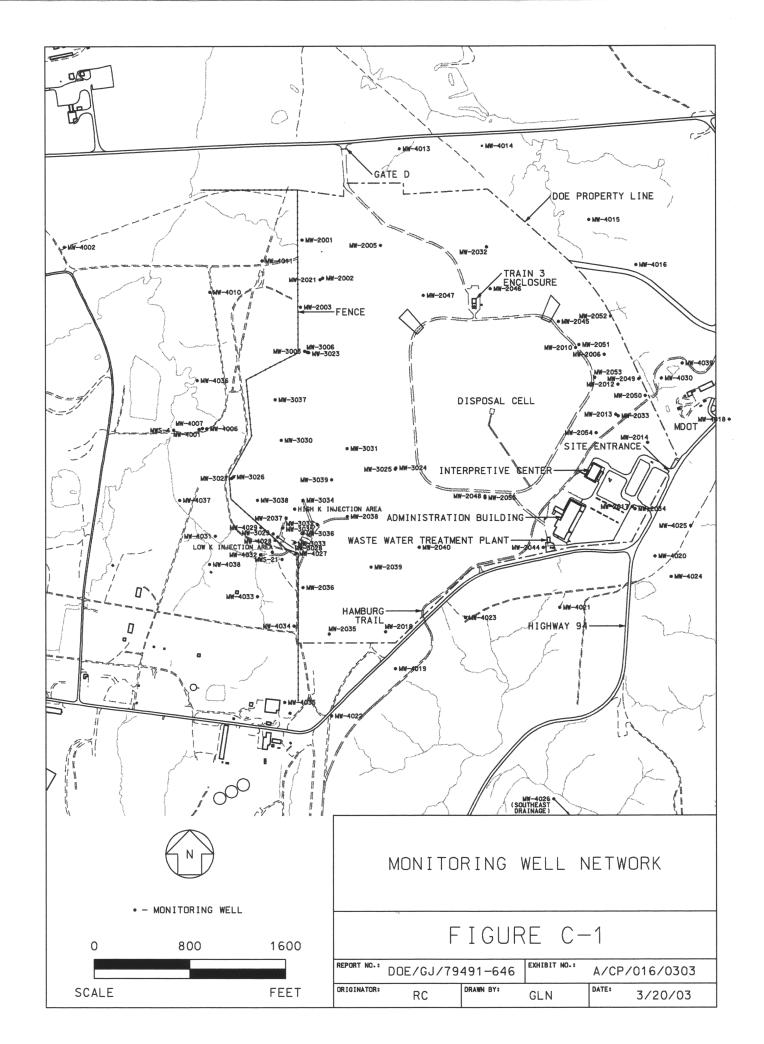
$$v = -Ki / n_e$$

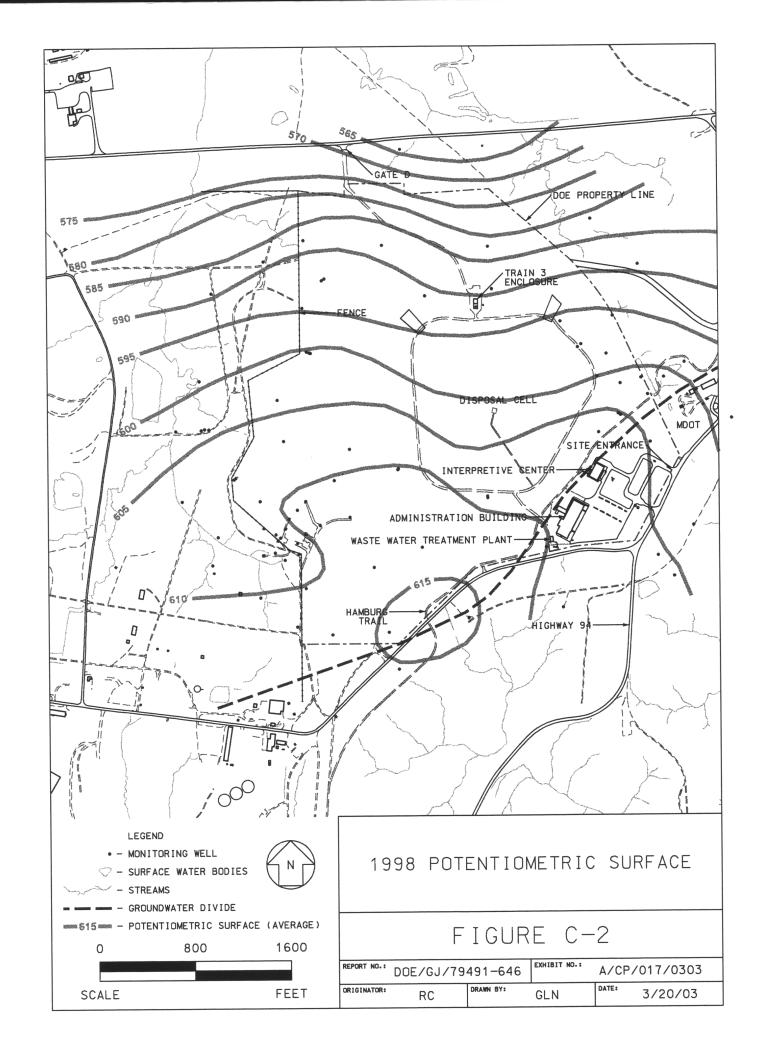
The average groundwater flow rate for each year is summarized in Table C-1.

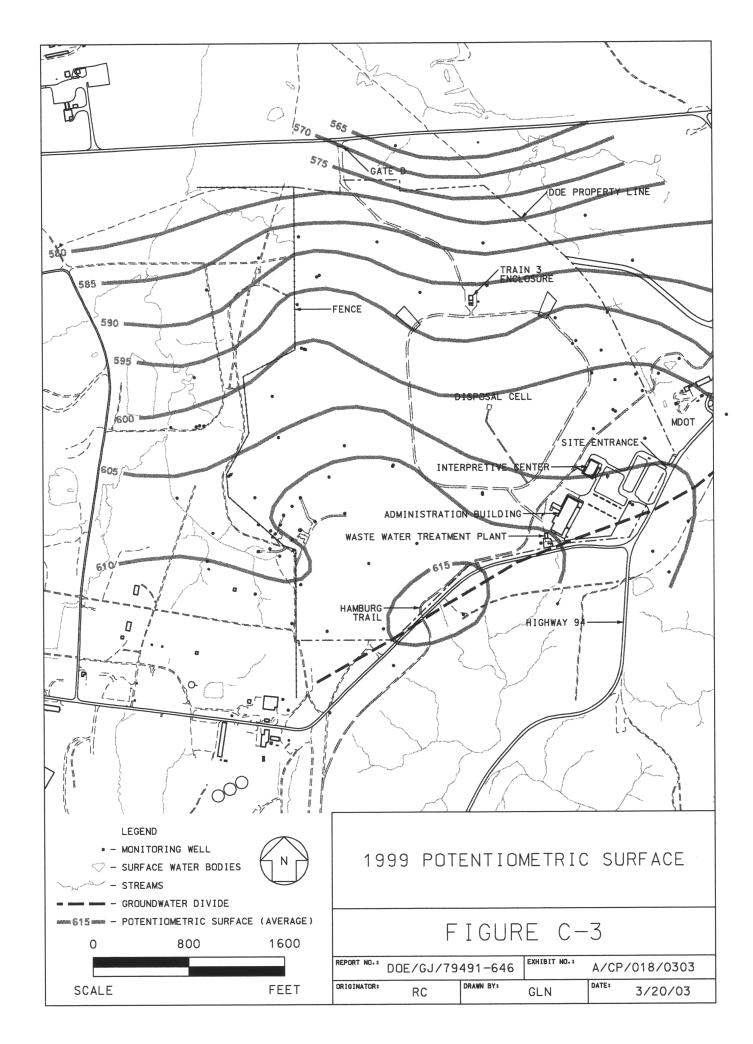
Table C-1 Average Groundwater Flow Rate From 1998 Through 2002

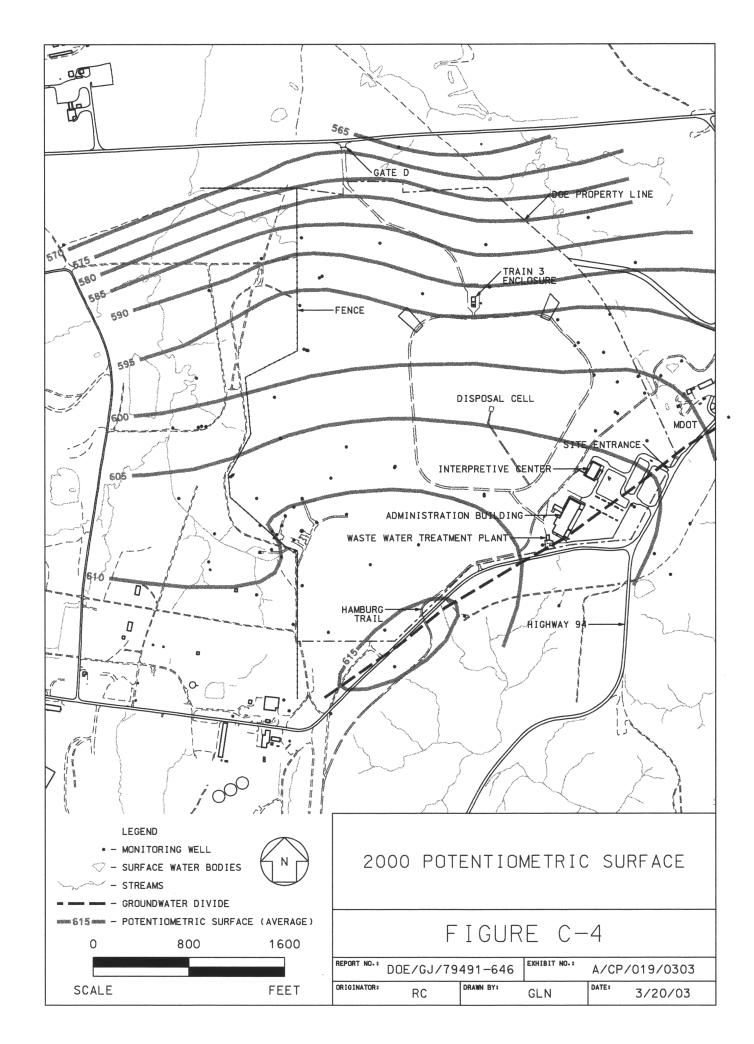
YEAR	Hydraulic	Effective Porosity ²	GW Elevation		Hydraulic Gradient	Average Flow Rate
	Conductivity (cm/s) 1		MW-2048 ³	MW-2032	(ft/ft) 4	(ft/day)
1998	(011110)	0.10	607.5	582.9	0.012	2.4
	_		607.5	583.0	0.012	2.4
1999	0.007		607.5	582.9	0.012	2.4
2000	0.007		607.3	582.9	0.012	2.4
2001			606.8	582.9	0.011	2.2
2002				302.9	0.011	

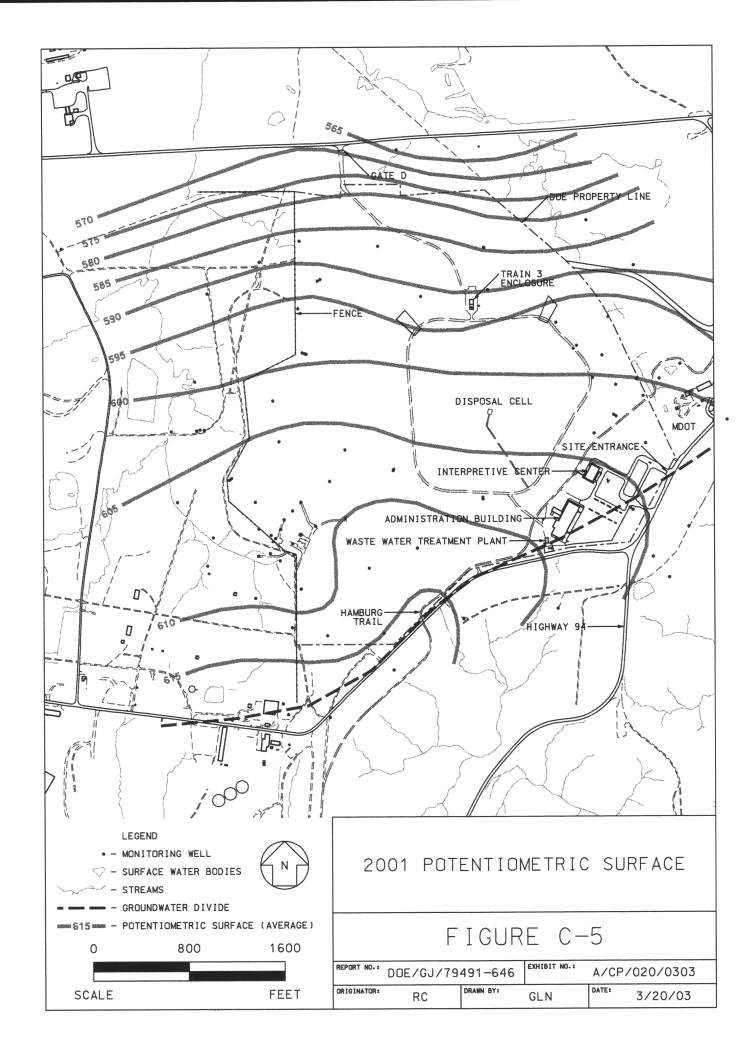
- 1 Average hydraulic conductivity using data from the cell monitoring wells.
- 2 Value selected to estimate maximum groundwater flow rate.
- 3 Groundwater elevation from MW-2055 was used for 2002.
- 4 Horizontal distance between MW-2032 and MW-2048 is 2,100 ft.

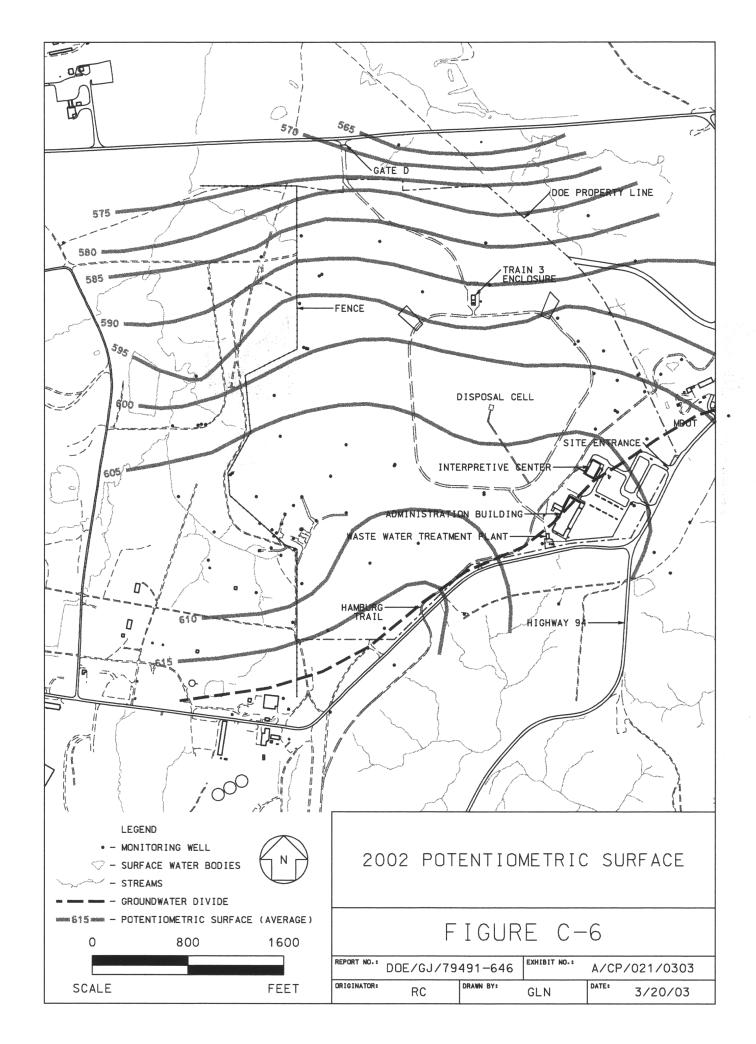












		YEAR				
LOCATION_	Data	1998	1999	2000	2001	2002
MW-2001	Average of WATER_ELEVATION	588.115	588.2175	587.9475	587.5625	588.0725
	Max of WATER_ELEVATION	588.73	588.58	588.48	587.69	588.43
	Min of WATER_ELEVATION	587.47	587.78	587.73	587.41	587.84
	Count of WATER_ELEVATION	4	4	4	4	4
MW-2002	Average of WATER_ELEVATION	593.214	592.9025	593.41	591.815	591.86
	Max of WATER ELEVATION	593.99	593.51	594.3	592.68	592.29
	Min of WATER ELEVATION	592.53	592.41	592.22	591.36	591.61
	Count of WATER ELEVATION	5	4	4	4	4
MW-2003	Average of WATER_ELEVATION	597.335	597.4875	597.0975	596.765	597.2525
	Max of WATER ELEVATION	597.63	597.73	597.6	596.99	597.79
	Min of WATER ELEVATION	596.98	597.19	596.33	596.47	596.79
	Count of WATER_ELEVATION	4	4	4	4	4
MW-2005	Average of WATER ELEVATION	587.424	587.6		587.233333	595.4425
	Max of WATER_ELEVATION	587.74	587.83	587.47	587.33	612.25
	Min of WATER_ELEVATION	587	587.44	587.38	587.16	589.79
	Count of WATER_ELEVATION	5	4	4	3	4
MW-2006	Average of WATER_ELEVATION	599.094	598.2475	597.62	598.545	598.05
2000	Max of WATER_ELEVATION	599.59	598.6	597.93	598.67	598.05
	Min of WATER_ELEVATION	598.67	598.01	597.22	598.29	598.05
	Count of WATER_ELEVATION	5	4	3	4	330.03
MW-2010	Average of WATER ELEVATION	599.565				
10100-2010	Max of WATER_ELEVATION	599.76				
	Min of WATER_ELEVATION	599.37				
	Count of WATER_ELEVATION	2				
MW-2012	Average of WATER_ELEVATION	602.936	601.0925	600.5925	600.9375	600.68
10100-2012	Max of WATER ELEVATION	604.2	601.28	600.72	601.07	600.81
	Min of WATER_ELEVATION	601.87	600.86	600.72	600.82	600.45
	Count of WATER_ELEVATION	5	4	4	4	000.43
MW-2013	Average of WATER_ELEVATION	604.518	603.8475	603.1675	604.31	603.3725
10100-2013	Max of WATER_ELEVATION	604.97	604.27	603.59	604.81	603.71
	Min of WATER_ELEVATION	604.15	603	602.73	604.03	602.72
	Count of WATER_ELEVATION	5	4	4	4	002.72
MW-2014	Average of WATER_ELEVATION	603.934	603.5825	603.2625	603.7825	603.705
10100-2014	Max of WATER_ELEVATION	604.2	603.84	603.55	604.05	603.705
	Min of WATER_ELEVATION	603.58	603.38	602.71	603.4	
	_				_	603.5
MM/ 2017	Count of WATER_ELEVATION	5	605.22	604.645	604.675	604.0405
MW-2017	Average of WATER_ELEVATION	605.15	605.23	604.615	604.675	604.8425
	Max of WATER_ELEVATION	605.36	605.54	604.84	605.3	605.07
	Min of WATER_ELEVATION	604.71	604.85	604.32	604.06	604.61
NAVA 0040	Count of WATER_ELEVATION	5	4	4	4	4
MW-2018	Average of WATER_ELEVATION	615.482				
	Max of WATER_ELEVATION	615.82				
	Min of WATER_ELEVATION	614.87				
10010010	Count of WATER_ELEVATION	5				
MW-2019	Average of WATER_ELEVATION	591.878				
	Max of WATER_ELEVATION	592.26				
	Min of WATER_ELEVATION	591.47				
	Count of WATER_ELEVATION	5				
MW-2021	Average of WATER_ELEVATION	589.256	589.1225	589.3025	588.27	588.6675
	Max of WATER_ELEVATION	589.98	589.66	591.44	588.5	589.21
	Min of WATER_ELEVATION	588.42	588.58	588.48	587.98	588.29
	Count of WATER_ELEVATION	5	4	4	4	4

1444 0000	TA	T 505.07	585.31	584.99	590.95	586.145
MW-2022	Average of WATER_ELEVATION	585.07	585.65	585.14	603.07	586.44
	Max of WATER_ELEVATION	585.67		584.71	584.74	586.01
	Min of WATER_ELEVATION	584.58	585.07	304.71	304.74	300.01
1414 0000	Count of WATER_ELEVATION	500.064	502.10	583.075	583.1025	583.19
MW-2023	Average of WATER_ELEVATION	582.964	583.18		583.2	583.34
	Max of WATER_ELEVATION	583.19	583.33	583.18		
	Min of WATER_ELEVATION	582.4	583.11	582.96	582.99	583.08
	Count of WATER_ELEVATION	5	507.055	500.07	4	F07 207F
MW-2024	Average of WATER_ELEVATION	566.956	567.255	566.67	566.6925	567.2875
	Max of WATER_ELEVATION	567.66	567.87	567	566.94	567.76
	Min of WATER_ELEVATION	566.2	566.88	566.49	566.16	566.83
	Count of WATER_ELEVATION	5	4	4	4	4
MW-2026	Average of WATER_ELEVATION	590.855				
	Max of WATER_ELEVATION	591.18				
	Min of WATER_ELEVATION	590.53				
	Count of WATER_ELEVATION	2				
MW-2027	Average of WATER_ELEVATION	591.95				
	Max of WATER_ELEVATION	592.24				
	Min of WATER_ELEVATION	591.66				
	Count of WATER_ELEVATION	2				
MW-2032	Average of WATER_ELEVATION	582.886	583.01	582.8675	582.09	582.92
	Max of WATER_ELEVATION	583.09	583.49	582.9	582.09	582.92
	Min of WATER_ELEVATION	582.79	582.74	582.82	582.09	582.92
	Count of WATER_ELEVATION	5	4	4	1	1
MW-2033	Average of WATER_ELEVATION	605.096	604.645	604.52	605.08	604.34
	Max of WATER_ELEVATION	605.55	604.85	604.71	605.27	604.69
	Min of WATER_ELEVATION	604.69	604.32	604.27	604.82	603.44
	Count of WATER_ELEVATION	5	4	4	4	4
MW-2034	Average of WATER_ELEVATION	606.302	606.495	605.61	605.63	605.6725
	Max of WATER_ELEVATION	606.81	606.71	605.68	606.6	606.11
	Min of WATER_ELEVATION	605.69	606.02	605.41	604.77	605.13
	Count of WATER_ELEVATION	5	4	4	4	4
MW-2035	Average of WATER_ELEVATION	613.708	613.7275		612.636667	615.7125
	Max of WATER_ELEVATION	614.08	613.97	613.36	612.91	615.79
	Min of WATER_ELEVATION	612.92	613.45	613.19	612.49	615.59
	Count of WATER_ELEVATION	5	4	4	3	4
MW-2036	Average of WATER_ELEVATION	610.552	610.755	609.9875	608.7975	609.4725
	Max of WATER_ELEVATION	611.57	610.91	610.21	610.21	609.61
	Min of WATER_ELEVATION	609.57	610.56	609.42	607.43	609.22
	Count of WATER_ELEVATION	5	4	4	4	4
MW-2037	Average of WATER_ELEVATION	610.422	610.6875	610.19	609.42	609.51
	Max of WATER_ELEVATION	612.21	610.8	610.38	609.86	609.78
	Min of WATER_ELEVATION	608.62	610.53	610.09	608.98	609.17
	Count of WATER_ELEVATION	5	4	4	2	4
MW-2038	Average of WATER_ELEVATION	611.146	612.125	613.025	610.0675	609.9025
	Max of WATER_ELEVATION	611.87	614.65	619.63	611.1	610.06
	Min of WATER_ELEVATION	610.45	611.01	610.66	609.02	609.81
	Count of WATER ELEVATION	5	4	4	4	4
MW-2039	Average of WATER_ELEVATION	612.602	611.8925	612.45	615.13	611.24
	Max of WATER ELEVATION	613.05	612.86	612.45	615.13	611.39
	Min of WATER_ELEVATION	611.86	609.6	612.45	615.13	611.05
	Count of WATER_ELEVATION	1	· · · · -			· · · · ·

					14	NAVA (00 15
611.065		611.933333		612.508	Average of WATER_ELEVATION	MW-2040
611.32		612.18	613.04	612.76	Max of WATER_ELEVATION	
610.91	610.74	611.71	612.31	612.21	Min of WATER_ELEVATION	
	4	3	4	5	Count of WATER_ELEVATION	
		616.06	612.25	612.384	Average of WATER_ELEVATION	MW-2041
		616.06	612.43	612.69	Max of WATER_ELEVATION	
		616.06	611.97	612.21	Min of WATER_ELEVATION	
		1	4	5	Count of WATER_ELEVATION	
		613.28	613.4975	613.822	Average of WATER_ELEVATION	MW-2042
		613.28	613.97	614.12	Max of WATER_ELEVATION	
		613.28	613.09	613.58	Min of WATER_ELEVATION	
		1	4	5	Count of WATER_ELEVATION	
		613.09	613.2925	611.088	Average of WATER_ELEVATION	MW-2043
		613.09	613.69	613.8	Max of WATER_ELEVATION	
		613.09	612.98	601.9	Min of WATER_ELEVATION	
		1	4	5	Count of WATER_ELEVATION	
				613.805	Average of WATER_ELEVATION	MW-2044
				613.84	Max of WATER_ELEVATION	
				613.77	Min of WATER_ELEVATION	
				2	Count of WATER_ELEVATION	
596.65	599.355	596.48	596.8625	597.01	Average of WATER_ELEVATION	MW-2045
596.94	605.88	596.6	597	597.25	Max of WATER_ELEVATION	
596.44	597.08	596.4	596.64	596.5	Min of WATER_ELEVATION	
4	4	4	4	5	Count of WATER_ELEVATION	
589.01	589.086667	589.105	589	588.842	Average of WATER_ELEVATION	MW-2046
589.06	589.18	589.22	589.08	589.21	Max of WATER_ELEVATION	
588.93	589.02	589.02	588.92	588.25	Min of WATER_ELEVATION	
4	3	4	4	5	Count of WATER_ELEVATION	
591.01	590.9275	595.643333	591.045	590.878	Average of WATER_ELEVATION	MW-2047
591.08	590.97	604.91	591.11	591.21	Max of WATER_ELEVATION	
590.92	590.89	590.97	590.93	590.43	Min of WATER_ELEVATION	
4	4	3	4	5	Count of WATER_ELEVATION	
	607.343333	607.485	607.51	607.54	Average of WATER_ELEVATION	MW-2048
	607.39	607.64	607.66	607.72	Max of WATER_ELEVATION	
	607.3	607.42	607.43	607.08	Min of WATER_ELEVATION	
	3	4	4	5	Count of WATER_ELEVATION	
599.365	599.5225				Average of WATER_ELEVATION	MW-2049
599.47	599.6				Max of WATER_ELEVATION	
599.23	599.43				Min of WATER_ELEVATION	
4	4				Count of WATER_ELEVATION	
601.03	601.1825				Average of WATER ELEVATION	MW-2050
601.13	601.35				Max of WATER ELEVATION	
600.9	600.99				Min of WATER_ELEVATION	
4	4				Count of WATER_ELEVATION	
598.88	598.94				Average of WATER_ELEVATION	MW-2051
598.99	598.95				Max of WATER_ELEVATION	
598.79						
JJU.13						
606.795						MW-2055
606.793						
606.66						
906.66 4						
_	598.93 2				Min of WATER_ELEVATION Count of WATER_ELEVATION Average of WATER_ELEVATION Max of WATER_ELEVATION Min of WATER_ELEVATION Count of WATER_ELEVATION	MW-2055

NAVA / 2002	A	E00 E17E	E00 447E	E00.0E	E07 E07E	E07.70
MW-3003	Average of WATER_ELEVATION	598.5175		598.05	597.5275	597.72
	Max of WATER_ELEVATION	599.22	598.85	598.11	597.86	597.95 597.43
	Min of WATER_ELEVATION	597.88	598.01	597.96	597.17	597.43
MM/ 2000	Count of WATER_ELEVATION	500 5075	500.57	500.0475	504.4005	504.77
MW-3006	Average of WATER_ELEVATION	592.5375	592.57	592.0475	591.1925	591.77
	Max of WATER_ELEVATION	593.5	593.08	592.74	591.45	592.35
	Min of WATER_ELEVATION	591.85	591.94	591.49	590.78	591.22
	Count of WATER_ELEVATION	4	4	4	4	4
MW-3019B	Average of WATER_ELEVATION	606.37				
	Max of WATER_ELEVATION	606.68				
	Min of WATER_ELEVATION	606.06				
	Count of WATER_ELEVATION	2				
MW-3023	Average of WATER_ELEVATION	600.935	601.17	598.3225	599.4375	599.8675
	Max of WATER_ELEVATION	601.74	601.27	600.34	599.96	599.9
	Min of WATER_ELEVATION	600.19	601.07	592.58	598.23	599.83
	Count of WATER_ELEVATION	4	4	4	4	4
MW-3024	Average of WATER_ELEVATION	601.28	601.74	599.9	600.5425	601.115
	Max of WATER_ELEVATION	601.64	602.38	600.77	600.95	601.65
	Min of WATER_ELEVATION	600.84	601.35	597.49	600.24	600.84
	Count of WATER_ELEVATION	4	3	4	4	4
MW-3025	Average of WATER_ELEVATION	610.1375	610.193333	609.235	608.1475	607.865
	Max of WATER_ELEVATION	610.4	610.33	609.52	608.78	608.06
	Min of WATER_ELEVATION	609.48	609.99	608.88	607.33	607.61
	Count of WATER_ELEVATION	4	3	4	4	4
MW-3026	Average of WATER_ELEVATION	607.3675	606.7475	605.8725	605.26	606.0375
	Max of WATER_ELEVATION	608.49	607.22	606.27	605.36	606.53
	Min of WATER_ELEVATION	606.14	606.39	605.05	605.14	605.67
	Count of WATER_ELEVATION	4	4	4	4	4
MW-3027	Average of WATER_ELEVATION	608.49	608.1425	607.9625	606.6825	607.005
	Max of WATER_ELEVATION	609.59	608.46	609.46	607.41	607.31
	Min of WATER_ELEVATION	607.66	607.89	607.31	606.36	606.7
	Count of WATER_ELEVATION	4	4	4	4	4
MW-3028	Average of WATER_ELEVATION	609.27	610.495	610.055	602.2625	609.2775
	Max of WATER_ELEVATION	610.17	610.62	610.19	609.72	609.45
	Min of WATER_ELEVATION	608.42	610.33	609.95	593.11	608.98
	Count of WATER_ELEVATION	3	4	4	4	4
MW-3029	Average of WATER_ELEVATION	609.146667	610.3725	609.9	607.515	609.15
	Max of WATER_ELEVATION	610.04	610.5	610.04	609.59	609.31
	Min of WATER_ELEVATION	608.3	610.21	609.77	604.89	608.84
	Count of WATER_ELEVATION	3	4	4	4	4
MW-3030	Average of WATER_ELEVATION				607.103333	607.425
	Max of WATER_ELEVATION				607.29	607.55
	Min of WATER ELEVATION				606.87	607.3
	Count of WATER_ELEVATION				3	4
MW-3031	Average of WATER_ELEVATION				608.76	608.6025
	Max of WATER_ELEVATION				608.93	609.01
	Min of WATER ELEVATION				608.51	608.01
	Count of WATER ELEVATION				3	10.00
MW-3032	Average of WATER_ELEVATION				608.93	609.3525
11111 0002	Max of WATER_ELEVATION				608.93	
	Min of WATER_ELEVATION				608.93	609.51
	Count of WATER_ELEVATION					609.11
	TOURIST MATER_ELEVATION	1			1	4

MW-3033	Average of MATER ELEVATION	1		· · · · · · · · · · · · · · · · · · ·	E0E 206667	
10100-3033	Average of WATER_ELEVATION				585.386667	
	Max of WATER_ELEVATION				587.7	
	Min of WATER_ELEVATION				582.38	
NAVA / 2024	Count of WATER_ELEVATION				3	
MW-3034	Average of WATER_ELEVATION				608.313333	
	Max of WATER_ELEVATION				610.1	
	Min of WATER_ELEVATION				606.07	
	Count of WATER_ELEVATION				3	
MW-3035	Average of WATER_ELEVATION				607.773333	
	Max of WATER_ELEVATION				608.97	
	Min of WATER_ELEVATION				605.59	
	Count of WATER_ELEVATION				3	-
MW-3036	Average of WATER_ELEVATION				607.976667	
	Max of WATER_ELEVATION				609.61	
	Min of WATER_ELEVATION				605.54	609.01
	Count of WATER_ELEVATION				3	4
MW-3037	Average of WATER_ELEVATION					601.956667
	Max of WATER_ELEVATION					602.09
	Min of WATER_ELEVATION					601.78
	Count of WATER_ELEVATION					3
MW-3038	Average of WATER_ELEVATION					609.303333
	Max of WATER_ELEVATION					609.4
	Min of WATER_ELEVATION					609.16
	Count of WATER_ELEVATION					3
MW-3039	Average of WATER_ELEVATION					608.82
	Max of WATER_ELEVATION					608.89
	Min of WATER_ELEVATION					608.76
	Count of WATER_ELEVATION					3
MW-4001	Average of WATER_ELEVATION	602.814	602.1575	601.5775	601.7725	602.48
	Max of WATER_ELEVATION	603.71	602.72	601.89	602.85	603.83
	Min of WATER ELEVATION	601.8	601.54	601.28	601.1	601.87
	Count of WATER_ELEVATION	5	4	4	4	4
MW-4002	Average of WATER ELEVATION	575.842	570.5225	568.86	570.436667	577.0675
	Max of WATER_ELEVATION	584.9	576.37	569.15	573.8	
	Min of WATER ELEVATION	568.16	568.27	568.71	568.75	568.8
	Count of WATER_ELEVATION	5	4	4	3	4
MW-4003	Average of WATER_ELEVATION	614.318	615.08			
	Max of WATER ELEVATION	615.28	615.08			
	Min of WATER_ELEVATION	612.61	615.08			
	Count of WATER_ELEVATION	5	1			
MW-4004	Average of WATER_ELEVATION	612.094	612.12			
	Max of WATER ELEVATION	612.66	612.12			
	Min of WATER ELEVATION	611.56	612.12			
	Count of WATER_ELEVATION	5	1			
MW-4005	Average of WATER ELEVATION	610.19	610.31	 		
	Max of WATER_ELEVATION	610.96	610.31			
	Min of WATER_ELEVATION	609.32	610.31			
	Count of WATER_ELEVATION	5	1			
MW-4006	Average of WATER_ELEVATION	602.938	602.1175	601.61	601 5005	600.00
19199 7000	Max of WATER_ELEVATION	1			601.5025	602.26
	Min of WATER_ELEVATION	603.74 602.27	602.46	601.84	601.75	603.33
	Count of WATER_ELEVATION		601.64	601.4	601.14	601.66
	TOURIST WATER_ELEVATION	5	4	4	4	4

MW-4007	Average of WATER ELEVATION	596.358	596.0175	594.8825	594.4625	595.325
IVIVV-4UU/	Average of WATER_ELEVATION Max of WATER_ELEVATION	596.358	596.0175	595.01	594.4625	596.38
	Min of WATER ELEVATION	595.68	595.15	594.65	593.93	594.48
	Count of WATER_ELEVATION	5	393.13	354.03	393.93	394.40
MW-4008	Average of WATER ELEVATION	597.67	598.1		4	
10100-4006	Max of WATER ELEVATION	598.38	598.1			
	Min of WATER_ELEVATION	597.01	598.1			
	<u> </u>	5	1			
NAVA 4000	Count of WATER_ELEVATION	594.2	594.73			
MW-4009	Average of WATER_ELEVATION					
	Max of WATER_ELEVATION	595.45	594.73			
	Min of WATER_ELEVATION	593.21	594.73			
1010	Count of WATER_ELEVATION	5	500.54	-		
MW-4010	Average of WATER_ELEVATION	589.39	590.54			
	Max of WATER_ELEVATION	591.04	590.54			
	Min of WATER_ELEVATION	588.02	590.54			
	Count of WATER_ELEVATION	5	1	504.0075		500.70
MW-4011	Average of WATER_ELEVATION	592.01	591.935	591.9275	591.5875	592.76
	Max of WATER_ELEVATION	592.79	592.45	592.93	591.86	593.3
	Min of WATER_ELEVATION	591.32	591.3	591.17	591.27	591.76
	Count of WATER_ELEVATION	5	4	4	4	4
MW-4012	Average of WATER_ELEVATION	570.8725	571.08			
	Max of WATER_ELEVATION	571.25	571.08			
	Min of WATER_ELEVATION	570.72	571.08			
	Count of WATER_ELEVATION	4	1		,	
MW-4013	Average of WATER_ELEVATION	560.2375	560.2525	561.0325	561.0625	560.4125
	Max of WATER_ELEVATION	560.38	560.33	563.2	563.35	560.5
	Min of WATER_ELEVATION	560.16	560.22	560.22	560.29	560.36
	Count of WATER_ELEVATION	4	4	4	4	4
MW-4014	Average of WATER_ELEVATION	560.4875	560.6375	560.5975	560.9825	561.23
	Max of WATER_ELEVATION	561.02	561.06	561.05	561.1	561.66
	Min of WATER_ELEVATION	560.21	560.33	560.07	560.8	561.04
	Count of WATER_ELEVATION	4	4	4	4	4
MW-4015	Average of WATER_ELEVATION	581.39	581.47	580.8175	581.74	581.9575
	Max of WATER_ELEVATION	582.51	582.53	580.98	581.98	583.01
	Min of WATER_ELEVATION	580.65	580.87	580.64	581.51	581.34
	Count of WATER_ELEVATION	4	4	4	4	4
MW-4016	Average of WATER_ELEVATION	589.14	589.25			
	Max of WATER_ELEVATION	589.42	589.25			
	Min of WATER_ELEVATION	588.82	589.25			
	Count of WATER_ELEVATION	4	1			
MW-4018	Average of WATER_ELEVATION	598.095	598.08			
	Max of WATER_ELEVATION	598.56	598.08			
	Min of WATER_ELEVATION	597.46	598.08			
	Count of WATER_ELEVATION	4	1			
MW-4019	Average of WATER_ELEVATION	613.105				
	Max of WATER_ELEVATION	613.26				
	Min of WATER_ELEVATION	612.95				
	Count of WATER_ELEVATION	2				
MW-4020	Average of WATER_ELEVATION	604.9875	605.26	604.2475	604.06	604.6225
	Max of WATER_ELEVATION	605.3	605.73	604.43	604.32	604.89
	Min of WATER_ELEVATION	604.4	604.83	604.02	603.59	604.24
	Count of WATER_ELEVATION	4	4	4	3	4
	10001101	1	7	7	3	-

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Count of WATER_ELEVATION	4 5775 	4 614.5025 615.07 613.93 4
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Max of WATER_ELEVATION 610.23 610.64 610.22 6 Min of WATER_ELEVATION 608.41 610.42 609.93 6	09.38	609.2825
Min of WATER_ELEVATION 608.41 610.42 609.93 6	09.97	609.41
l	08.79	609.03
Count of WATER ELEVATION 3 4 4	2	4
	.6625	609.34
	09.72	609.73
	05.22	608.98
Count of WATER_ELEVATION 4 4 4	4	4
	.0775	609.2475
	609.7	609.42
	05.85	608.95
Count of WATER_ELEVATION 4 4 4	4	4
	.1625	597.21
Max of WATER_ELEVATION 5	97.29	597.36
Min of WATER_ELEVATION 5	97.06	597.13
Count of WATER_ELEVATION	4	4
MW-4031 Average of WATER_ELEVATION 608.36	6667	609.3725
Max of WATER_ELEVATION 6	08.66	609.61
Min of WATER_ELEVATION 6	08.03	609.24
Count of WATER_ELEVATION	3	4
MW-4032 Average of WATER_ELEVATION 6	08.54	609.5
Max of WATER_ELEVATION 6	88.80	609.72
Min of WATER_ELEVATION	608.2	609.25
Count of WATER_ELEVATION	2	4
MW-4033 Average of WATER_ELEVATION	608.5	609.3
		000 40
	08.72	609.42
Count of WATER_ELEVATION	08.72 08.28	609.42

MW-4034	Average of WATER ELEVATION				610.71	611.1525
	Max of WATER ELEVATION				610.78	611.21
	Min of WATER ELEVATION				610.64	611.02
	Count of WATER ELEVATION				2	4
MW-4035	Average of WATER ELEVATION				623.24	623.8925
	Max of WATER ELEVATION				623.42	624.08
	Min of WATER ELEVATION				623.06	623.43
	Count of WATER ELEVATION				2	4
MW-4036	Average of WATER_ELEVATION					592.51
	Max of WATER_ELEVATION					593.23
	Min of WATER_ELEVATION					591.63
	Count of WATER_ELEVATION					4
MW-4037	Average of WATER_ELEVATION					608.2525
	Max of WATER_ELEVATION					609.61
	Min of WATER_ELEVATION					607.19
	Count of WATER_ELEVATION					4
MW-4038	Average of WATER_ELEVATION					609.51
	Max of WATER_ELEVATION					609.64
	Min of WATER_ELEVATION					609.23
	Count of WATER_ELEVATION					4
MWS-21	Average of WATER_ELEVATION	609.33875	606.25	609.89	608.44	609.4075
	Max of WATER_ELEVATION	610.28	610.47	609.89	610.28	609.68
	Min of WATER_ELEVATION	608.93	602.03	609.89	606.6	609.19
	Count of WATER_ELEVATION	8	2	1	4	4
MWS-3	Average of WATER_ELEVATION	594.303333	598.915			
	Max of WATER_ELEVATION	594.73	603.11			
	Min of WATER_ELEVATION	593.97	594.72			
	Count of WATER_ELEVATION	3	2			
MWS-4	Average of WATER_ELEVATION	602.581429	602.675	601.39	601.556667	602.41
	Max of WATER_ELEVATION	602.94	602.85	601.39	601.94	603.53
	Min of WATER_ELEVATION	601.62	602.5	601.39	601.12	601.84
l	Count of WATER_ELEVATION	7	2	1	3	4

Appendix L Well Field Contingency Plan

L1.0 Planning and Preparation

Under this contingency plan, which supersedes the *Well Field Contingency Plan* (DOE 1992b), any production capacity lost to the existing well field due to confirmed contaminant migration from the Weldon Spring quarry will be replaced. While it is highly unlikely that such measures will be implemented, this plan defines the minimum planning and preparation required to facilitate a rapid and effective response. Planning and preparation measures include the following:

- Selection of a reliable alternate source of water to replace or supplement the existing well field
- Preparation of a plan for data collection to facilitate development of the selected alternate source.
- Development of design criteria for use in design and construction of the alternate source infrastructure.

L1.1 Selection of Alternate Source

Criteria and alternatives for contingency planning were developed using modified value engineering principles. Modified value engineering is an alternative evaluation process that parallels the CERCLA philosophy of remedial alternative development that is not based upon cost unless all other criteria (i.e., effectiveness, implementability, etc.) are equal. This process was performed as outlined in *Alternative Evaluation Study Manual* (DOE 2000).

Two broad potential scenarios were considered as part of alternative evaluation: (1) a portion of the well field is threatened, requiring partial replacement of the water supply; and (2) the entire well field is threatened, requiring replacement of the entire water supply from the existing well field.

The criteria used to evaluate the alternatives were effectiveness, technical feasibility, degree of disruption, public acceptance, regulatory requirements, cost, and impact on the present treatment system. By applying these criteria, all but the top three alternatives for each scenario were quickly eliminated (Table L-1). Further evaluation of the remaining alternatives led to the selection of a proposed alternative. The evaluation and selection process is described in the report *St. Charles County Well Field Summary of Alternatives for Contingency Plans* (DOE 1992a).

Table L-1. Alternatives Considered for Water Supply Replacement Scenarios

	Rank of Alternative		
Alternative	Partial Replacement Scenario	Full Replacement Scenario	
New well(s) in existing well field	2	7	
New well(s) in Darst Bottoms upstream of existing well field	1	1	
Modify existing well system	10	10	
Change pumping scheme of existing wells	6	9	
Utilize existing pipeline from St. Louis	5	11	
New pipeline to Howard Bend Plant	4	3	
Treat Missouri River surface water	3	2	
Find bedrock source of water at another site	7	6	
Treat and use contaminated water	11	8	
Protect well field with a slurry wall	8	4	
Redirection of existing capacities	9	5	
No action	Not appropriate	Not appropriate	

The selected alternative is the installation of additional water supply wells in the Darst Bottoms to the south of the present well field (Figure L-1). Although this location is within the same aquifer as the present well field, the replacement location is upgradient of the contaminant source, the Weldon Spring quarry. Hence, given that action levels for contaminants are conservative (low), the replacement well field location would be unaffected by contaminant migration either from the quarry or a potentially tainted well field to the north.

L1.2 Preparation of a Plan for Hydrogeologic Investigation

A plan will be prepared for a hydrogeologic investigation required to obtain the information necessary to develop the alternate source of ground water. This plan will identify the activities, sampling, and testing required to assess the hydrogeologic characteristics of the replacement well field area. While the hydrogeologic characteristics of the replacement well field location are probably quite similar to the present well field, additional data and testing will be required to ensure an adequate assessment, and to ensure that engineering design is optimized to meet production needs.

L1.3 Design Criteria

Engineering design criteria will be established for use in design and construction of the alternate water supply. Design criteria will address:

- Functional requirements relative to interface with the existing well field and treatment plant.
- Performance requirements relative to production capacity.
- Phased response (requirements for partial versus full replacement).
- Water quality requirements.
- Well sitting and construction.

In the event an alternate source of drinking water is required, engineering design and construction shall proceed based on the design criteria established under this plan.

L1.4 Access

Should the need arise, access for data collection purposes, well installation, and pipe line construction will be coordinated with the affected private landowners and St. Charles County officials. As an interim measure, private landowners who would be affected by construction of a replacement well field were contacted by a U.S. Department of Energy (DOE) representative who explained the contingency plan and outlined the potential for a request for access to be made at some future time.

L1.5 Installation of Replacement Wells

In the event that contaminants from the Weldon Spring quarry are detected above action levels established under this plan, the following steps will be taken to install a replacement well field:

- Access will be obtained from affected landowners.
- Subcontractor services will be procured for drilling of production and test wells and acquisition of other data prescribed as part of the hydrogeologic investigation.
- Field activities will be initiated as detailed in the hydrogeologic investigation plan.
- Design of components necessary to perform drilling, install wells, pumps, and piping, and construct pumping facilities and controls will be accelerated.
- Procurement of materials will be accelerated for pumps, piping, casing, screens, and all
 appurtenances required to complete construction of the replacement well field to production
 standards.
- The replacement well field will be installed under the direction of DOE.

L1.6 Permits

Construction permits would be required from the MDNR and St. Charles County as well as a permit from the Darst Bottoms Levee District in order to install the replacement wells. The permit process is estimated to take between 60 and 90 days (DOE 1992b).

L1.7 Schedule

Assuming that construction would proceed on several tasks simultaneously, it is estimated that a minimum of 2 months will be required for construction after permits are obtained. Allowing 60 days for engineering and the preparation of permit applications, about 200 days would be required from the start of engineering through the start up of the pumps (DOE 1992b). The estimated implementation schedule is illustrated in Figure L–2.

During the period of time required to complete installation of the replacement well field, the present well field would operate without the reserve provided by the affected wells. In a worst case scenario, the present well field might not meet production demands during the period of new well field construction. In this instance, service demands for St. Charles County Plant No. 1 would have to be met through an alternate source or rationing (such as water used for lawn care and car washing, etc.) until the replacement well field went on line or demand subsided due to the normal demand cycle.

L1.8 Well Design

Figure L-3 illustrates the preliminary design of the replacement wells.

L2.0 References

U.S. Department of Energy (DOE), 1992a. *St. Charles County Well Field Summary of Alternatives for Contingency Plans*, DOE/OR/21548-285, prepared by L.G. Zambrana Consultants, Inc. and Woodward-Clyde Consultants for the U.S. Department of Energy Oak Ridge Operations Office, Weldon Spring Site Remedial Action Project, Weldon Spring, Missouri, May.

-----, 1992b. *Well Field Contingency Plan*, DOE/OR/21548-340, U.S. Department of Energy Oak Ridge Operations Office, Weldon Spring Site Remedial Action Project, Weldon Spring, Missouri, November.

-----, 2000. *Alternative Evaluation Study Manual*, Rev. 1, DOE/OR/21548-640, U.S. Department of Energy Oak Ridge Operations Office, Weldon Spring Site Remedial Action Project, Weldon Spring, Missouri, January.

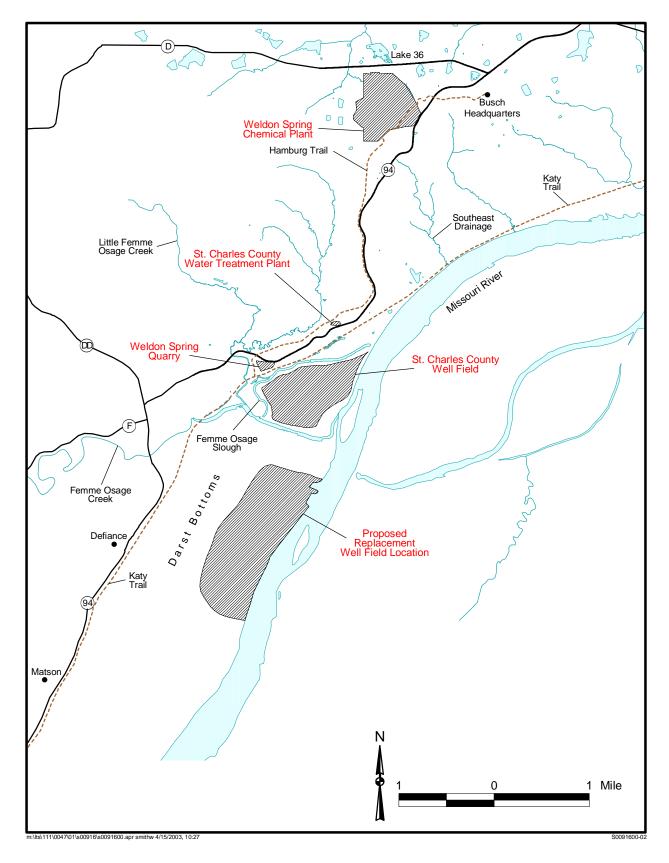


Figure L-1. Proposed Replacement Well Field Location

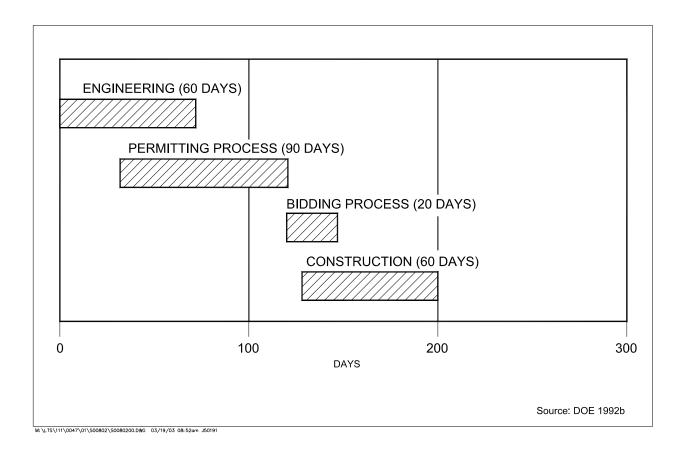


Figure L-2. Estimated Replacement Well field Installation Schedule

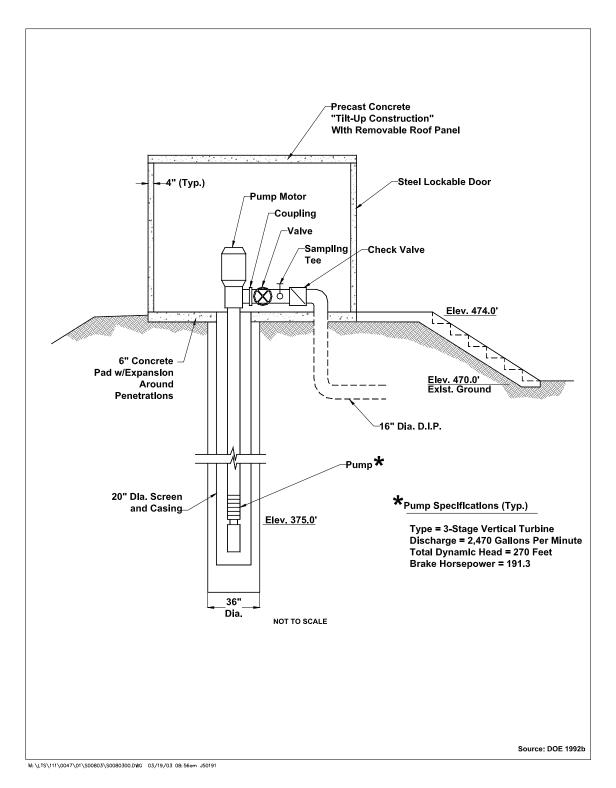


Figure L-3. Proposed Typical Replacement Well Schematic

End of current text